

## **Emissions of Ammonia from Light-Duty Vehicles**

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### **ABSTRACT**

Emissions of  $\text{NH}_3$  were measured with the FTP cycle for 39 vehicles ranging in technology from non-catalyst and Tier 0 vehicles to ULEV vehicles.  $\text{NH}_3$  measurements were performed using Fourier Transform Infrared spectroscopy (FTIR). Results showed that  $\text{NH}_3$  emissions averaged 54 mg/mi with a range from <4 to 177 mg/mi.  $\text{NH}_3$  emissions did not necessarily decline significantly like those of the regulated pollutants with progressive improvements in emission control systems. A subset of 5 vehicles was tested over the US06, NYCC, and a high-speed freeway cycle for comparison with the FTP cycle.  $\text{NH}_3$  emissions showed a strong cycle dependence, with increased emissions under more aggressive driving conditions, consistent with the formation of higher  $\text{NH}_3$  emissions under richer conditions. The onset of  $\text{NH}_3$  emissions typically occurred after catalyst light off, near when the catalyst reached its equilibrium temperature. Initial studies showed that  $\text{NH}_3$  emissions decreased as the sulfur content in the fuel was increased.

## 1.0 Introduction

As ambient air quality standards become increasingly stringent, the role of unregulated vehicle emissions is becoming more important. Ammonia ( $\text{NH}_3$ ) is one compound that has received attention recently. Studies of  $\text{NH}_3$  emissions from vehicle exhaust date back to the late 1970s.<sup>1-3</sup> More recently, several studies have indicated that  $\text{NH}_3$  emissions from vehicles may be greater than previous thought, although there is a range of estimates for  $\text{NH}_3$  emissions. These include studies in tunnels,<sup>4,5,6</sup> remote sensing studies,<sup>7</sup> using chassis dynamometers,<sup>8,9,10</sup> and using dedicated vehicles.<sup>11</sup>

Mobile sources are the third largest component of the  $\text{NH}_3$  emissions inventory in the South Coast Air Basin (SoCAB) that surrounds Los Angeles representing approximately 18% of the inventory.<sup>12</sup> Based on tunnel measurements made by Fraser and Cass, the emissions factors were found to be 33.2 tons/day in SoCAB.<sup>12</sup> For comparison, the emission factor for livestock, the largest single contributor, was found to be 60.37 tons/day.<sup>12</sup> From an air quality standpoint,  $\text{NH}_3$  is unregulated but it is an important precursor gas that can react in the atmosphere to produce particulate matter (PM), such as ammonium nitrate ( $\text{NH}_4\text{NO}_3$ ) or ammonium sulfates ( $\text{NH}_4\text{SO}_4$ ,  $\text{NH}_4(\text{SO}_4)_2$ ). Analysis of ambient PM indicates that ammonium composed from 14.0 to 17.0% of the total fine PM ( $\text{PM}_{2.5}$ ) at various locations in SoCAB.<sup>13</sup>

To better quantify the vehicle contribution to the  $\text{NH}_3$  inventory, it is important to have a robust database of  $\text{NH}_3$  emission rates from vehicles. The objective of this study was to measure the emission levels of  $\text{NH}_3$  from an in-use fleet of vehicles. For this project, 39 vehicles were tested over the FTP cycle. This included Low Emission Vehicles (LEVs) that will compose a larger portion of the in-use fleet over the next 5 to 10 years. A subset of 5 of these vehicles was also tested over the US06, New York City Cycle (NYCC), and a high speed freeway cycle for comparison. Additional experiments were also conducted to evaluate the repeatability of  $\text{NH}_3$  measurements and the impact of fuel sulfur levels on  $\text{NH}_3$  emissions. Measurements were conducted using Fourier Transform Infrared spectroscopy (FTIR), which can measure mass emission rates of compounds like  $\text{NH}_3$  in near real-time as shown by researchers at the Ford Motor Company and elsewhere for nearly two decades.<sup>14</sup> The results of this study are described in further detailed in the following paper.

## **2.0 Experimental Procedures**

### **2.1 Vehicle Recruitment**

The 39 test vehicles were recruited from several sources including private owners, the University of California at Riverside campus fleet, and rental car companies. A breakdown of the test vehicles by manufacturer is provided in Table 1 with a more complete vehicle description provided in Appendix A. Although the primary goal of the project was to measure NH<sub>3</sub> in conjunction with CE-CERT's on-going projects, the test matrix shows that a reasonable distribution of the major manufacturers was obtained. This study focused primarily on newer cars with all but 5 of the vehicles being 1990 and newer model years. For the 1990 and newer vehicles, the average age of the test fleet was 1996. The test matrix also included a range of different emissions control technology levels including 14 pre-Tier 1 vehicles, 11 Tier 1 vehicles, 8 TLEV vehicles, 1 NLEV vehicle, 2 LEV vehicles, and 3 ULEV vehicles.

### **2.2 Protocol for Vehicle Testing**

All vehicles were tested over one Federal Test Procedure (FTP) to obtain mass emission rates for total hydrocarbons (THC), non-methane hydrocarbons (NMHC), carbon monoxide (CO), nitrogen oxides (NO<sub>x</sub>), and NH<sub>3</sub>. Replicate FTPs were performed on 5 of these vehicles. A subset of 5 vehicles was also tested over the US06, New York City Cycle (NYCC), and a high speed freeway cycle. For two vehicles, some initial tests were also conducted to evaluate the potential impact of fuel sulfur levels on NH<sub>3</sub> emissions. The regulated pollutants were measured using the standard techniques as outlined in the Code of Federal Regulations for the FTP (CFR Part 86, Subpart B).

NH<sub>3</sub> emissions were collected using a Pierburg AMA/Mattson FTIR system. The FTIR samples from the dilution tunnel and makes one measurement every three seconds. The minimum detection limits for NH<sub>3</sub> were 4 mg/mi over the FTP cycle. The absorption cell for the FTIR has a volume of 5 liters and the residence time in the cell is approximately 10 seconds. To adjust the model emissions data to correct for this time response, a well-mixed flow cell model is used. The use of a well-mixed flow cell model for analysis of modal emissions data is described in greater detail by Truex et al.<sup>15</sup> A three-second data average is also applied to the data prior to using the well-mixed flow cell model.

All tests were conducted in CE-CERT's Vehicle Emission Research Laboratory (VERL) equipped with a Burke E. Porter 48-inch single-roll electric dynamometer. Sampling was conducted using VERL's 10 inch diameter dilution tunnel and CVS flow rates of 350 SCFM. Since ammonia is a relatively reactive compound, a heating pad maintained at a temperature of 250°F was wrapped around the transfer tube for some of the experiments to minimize the loss of ammonia through the sampling system. A comparison of tests run with and without the heating pad showed no difference in the observed NH<sub>3</sub> emission levels, however.

All but 5 vehicles were tested with the gasoline in the tank at the time the vehicle was procured for testing. Since the specifications for California Phase 2 gasoline are relatively narrow and must provide equivalent emissions under California's Predictive Model, any effects due to testing with in-tank fuel should be negligible. The other vehicles were tested on California Phase 2 certification fuel (2 vehicles) and industry average RFA gasoline (3 vehicles). For two vehicles, tests were also conducted at two different sulfur levels. These fuels were certification grade California Phase 2 gasoline with nominal sulfur levels of 30 and 330 ppmw. These fuels were obtained from Philips Petroleum Chemical Company in Borger, TX. The vehicles for the fuel sulfur tests were preconditioned using procedures used in previous Auto/Oil research programs.<sup>16</sup>

### **3.0 Emissions Test Results**

A summary of the FTP emission results is provided in Table 2 for the 39 vehicle test fleet. More detailed emissions results for these vehicles are provided in Appendix A. These results show that NH<sub>3</sub> emissions ranged from <4 to 177 mg/mi with an average of 54 mg/mi. These NH<sub>3</sub> emission levels are similar to those found in previous studies. Chassis dynamometer tests conducted on a fleet of 75 in-use Canadian and United States (US) vehicles over a hot 505 cycle showed a range in NH<sub>3</sub> emissions from <1 mg/mi to nearly 300 mg/mi.<sup>8</sup> The present results are also very comparable to results obtained in tunnel studies, being slightly lower than those found by Fraser and Cass (116 mg/mi)<sup>4</sup> and Kean et al. (79 mg/mi)<sup>5</sup> and slightly higher than those found by Gertler et al. (15.1 ± 4.3 mg/mi)<sup>6</sup>. Given the differences that can be found in the fleet composition and testing conditions for the various studies, these reported NH<sub>3</sub> emissions rates show relatively good comparability. Overall, the results are consistent with those of previous

studies showing that while NH<sub>3</sub> emissions from vehicles are below those of the regulated pollutants, they can still make an important contribution to the overall inventory.

Results of replicate NH<sub>3</sub> measurements are presented in Table 3. In general, the replicates indicate that NH<sub>3</sub> emissions are probably repeatable within 10 to 20%, although for one vehicle the variability appeared to be considerably greater.

A histogram of the NH<sub>3</sub> emissions is presented in Figure 1. These data show that NH<sub>3</sub> emissions for 12 of the 39 vehicles were 10 mg/mi or less. Above this level, there was considerable range in the emission levels, indicating that NH<sub>3</sub> emissions can vary significantly based on vehicle technology and control strategy. The individual vehicle data for some of the highest NH<sub>3</sub> emitting vehicles are presented in Table 4. These data show that, in some cases, NH<sub>3</sub> emissions can have emissions similar to those of other regulated pollutants such as THC and NO<sub>x</sub>.

Average emission results for THC, CO, NO<sub>x</sub>, and NH<sub>3</sub> are presented in Figure 2 as a function of vehicle certification category. These results show that the emission levels for regulated pollutants have decreased significantly over the years. NH<sub>3</sub> emissions, however, did not necessarily decline as significantly for the range of technology categories tested. As shown in Table 4, the highest NH<sub>3</sub> emitters came from range of vehicle technology categories including Tier 0, Tier 1 and TLEV. It is still important to note, however, that only a few vehicles were tested for the low emission vehicle technology categories and that more data would be needed to provide a more definite comparison of NH<sub>3</sub> emissions for different technology categories. It should also be noted that most of the LEV and ULEV vehicles did have NH<sub>3</sub> emissions near the detection limit.

NH<sub>3</sub> emissions for the US06, NYCC, and the high-speed freeway cycles are presented in Figure 3. Similar to other regulated pollutants, NH<sub>3</sub> emissions were found to increase significantly over the more aggressive driving cycles. This is true even for vehicles that have relatively low levels of NH<sub>3</sub> emissions over the FTP. The results are consistent with previous studies, which have shown that NH<sub>3</sub> emissions can increase significantly under rich operating conditions.<sup>3,11</sup>

NH<sub>3</sub> modal emissions plotted against vehicle speed are presented in Figure 4 for the FTP and in Figure 5 for the US06, NYCC, and the high-speed freeway cycle for one of the test vehicles. The modal emissions show the transient nature of the NH<sub>3</sub> emissions throughout the driving cycle.

The modal data show the onset of NH<sub>3</sub> emissions occurs after catalyst light off, consistent with the formation of NH<sub>3</sub> over the catalyst surface. Experiments conducted on a separate vehicle where the catalyst bed temperature was monitored indicate that NH<sub>3</sub> emissions tend to occur as the catalyst gets closer to its equilibrium temperature rather than during the initial portion of catalyst light off.

Initial tests were conducted on two vehicles with gasoline containing sulfur levels of 30 and 330 ppmw over the FTP and US06. The results of these tests are presented in Table 5. Although the data are limited, for each of the vehicle/cycle combinations, higher NH<sub>3</sub> emissions were observed for each test(s) with the lower fuel sulfur level. Since NH<sub>3</sub> is primarily formed over the catalyst, these results suggest that sulfur could inhibit NH<sub>3</sub> formation on the catalyst by poisoning reaction sites for NH<sub>3</sub> formation.<sup>17</sup> Early engine dynamometer and simulated exhaust gas experiments have shown that increasing SO<sub>2</sub> concentrations in the exhaust suppresses the formation of NH<sub>3</sub>.<sup>18,19</sup> Chassis dynamometer measurements on two vehicles, however, showed that decreasing fuel sulfur content resulted in lower NH<sub>3</sub> emissions for one vehicle and had little effect on NH<sub>3</sub> emissions for the second vehicle.<sup>9</sup> The effect of gasoline sulfur levels on NH<sub>3</sub> emissions will be investigated more extensively in two upcoming studies in our laboratory.<sup>20,21</sup>

## 4.0 Summary and Conclusions

For the present program, a total of 39 vehicles were tested over the FTP with emissions measured for NH<sub>3</sub> using an FTIR. A subset of 5 vehicles was also tested over the US06, NYCC, and high speed freeway cycle. The major results of this study are:

- NH<sub>3</sub> FTP emissions for the vehicles tested ranged from <4 to 177 mg/mi and averaged 54 mg/mi. Of the 39 test vehicles, 12 had emissions below 10 mg/mi while the emissions from the remaining vehicles varied significantly depending on the specific vehicle. NH<sub>3</sub> emissions were comparable to those observed in previous studies and indicate that vehicles could be an important source for NH<sub>3</sub> emissions.
- NH<sub>3</sub> emissions did not necessarily decline significantly like those of the regulated pollutants with progressive improvements in emission control systems.
- NH<sub>3</sub> emission levels increased significantly for the more aggressive US06, NYCC, and high speed freeway cycles. This is consistent with previous studies that have shown increases in NH<sub>3</sub> emissions during periods of richer operation.
- Modal emissions showed that the onset of NH<sub>3</sub> emissions typically occurred after catalyst light off and near when the catalyst reached its equilibrium temperature.
- Initial studies showed that NH<sub>3</sub> emissions decreased as the sulfur content in the fuel was increased over the FTP and US06 cycles.

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**Table 1 List of Test Vehicles by Manufacturer**

<b>Manufacturer</b>	<b>Passenger Car</b>	<b>LD Truck</b>
<b>GM</b>	<b>3</b>	<b>9</b>
<b>Ford</b>	<b>4</b>	<b>5</b>
<b>Chrysler</b>	<b>3</b>	<b>1</b>
<b>Honda</b>	<b>6</b>	<b>0</b>
<b>Toyota</b>	<b>2</b>	<b>2</b>
<b>Nissan</b>	<b>2</b>	<b>0</b>
<b>Other</b>	<b>1</b>	<b>1</b>

**Table 2. Average FTP Emission Results**

	<b>THC</b> <b>g/mi</b>	<b>CO</b> <b>g/mi</b>	<b>NO<sub>x</sub></b> <b>g/mi</b>	<b>NH<sub>3</sub></b> <b>g/mi</b>
<b>Average</b>	<b>0.488</b>	<b>7.259</b>	<b>0.590</b>	<b>0.054</b>
<b>Median</b>	<b>0.183</b>	<b>2.312</b>	<b>0.247</b>	<b>0.046</b>
<b>High</b>	<b>4.938</b>	<b>116.98</b>	<b>3.709</b>	<b>0.177</b>
<b>Low</b>	<b>0.033</b>	<b>0.372</b>	<b>0.058</b>	<b>&lt;0.004</b>

**Table 3. Replicate Ammonia FTP Emissions**

	Cycle	NH <sub>3</sub> Emissions (mg/mi)	
		Test 1	Test 2
1992 Tier 0 PC FFV	FTP	118	119
1992 Tier 0 PC FFV	US06	196	224
1993 Tier 0 PC	FTP	36	18
1989 Tier 0 Van	FTP	64	52
1989 Tier 0 PC	FTP	<MDL	<MDL

**Table 4. Individual Vehicle Data for the Highest NH<sub>3</sub> Emitting Vehicles**

	THC (mg/mi)	NO <sub>x</sub> (mg/mi)	NH <sub>3</sub> (mg/mi)
1996 TLEV SUV	103	535	177
1995 Tier 1 PC	284	235	157
1993 Tier 0 PC	943	1,394	155
1996 Tier 1 PC	153	123	144
1992 Tier 0 FFV PC	165	172	119
1997 Tier 1 Van	267	412	118
1991 Tier 0 PC	330	1812	111
1996 Tier 1 PC	300	258	109
1993 Tier 1 SUV	203	410	95
1998 TLEV LDT	103	233	88

**Table 5. Emissions of NH<sub>3</sub> and Regulated Components at 2 Sulfur Levels**

<b>Vehicle</b>	<b>Cycle</b>	<b>S Level</b>	<b>THC g/mi</b>	<b>NMHC g/mi</b>	<b>CO g/mi</b>	<b>NO<sub>x</sub> g/mi</b>	<b>NH<sub>3</sub> g/mi</b>
<b>1992 Tier 0 FFV PC</b>	<b>FTP</b>	<b>30</b>	<b>0.177</b>	<b>0.138</b>	<b>2.791</b>	<b>0.176</b>	<b>0.118</b>
	<b>FTP</b>	<b>30</b>	<b>0.152</b>	<b>0.119</b>	<b>2.364</b>	<b>0.167</b>	<b>0.119</b>
		<b>Average</b>	<b>0.165</b>	<b>0.129</b>	<b>2.578</b>	<b>0.172</b>	<b>0.119</b>
	<b>FTP</b>	<b>330</b>	<b>0.211</b>	<b>0.161</b>	<b>3.250</b>	<b>0.226</b>	<b>0.086</b>
	<b>US06</b>	<b>30</b>	<b>0.225</b>	<b>0.164</b>	<b>9.984</b>	<b>0.619</b>	<b>0.195</b>
	<b>US06</b>	<b>30</b>	<b>0.289</b>	<b>0.227</b>	<b>11.870</b>	<b>0.599</b>	<b>0.224</b>
		<b>Average</b>	<b>0.257</b>	<b>0.196</b>	<b>10.927</b>	<b>0.609</b>	<b>0.210</b>
	<b>US06</b>	<b>330</b>	<b>0.322</b>	<b>0.246</b>	<b>13.184</b>	<b>0.805</b>	<b>0.161</b>
<b>1997 TLEV PC</b>	<b>FTP</b>	<b>30</b>	<b>0.054</b>	<b>0.051</b>	<b>0.514</b>	<b>0.058</b>	<b>0.038</b>
	<b>FTP</b>	<b>330</b>	<b>0.061</b>	<b>0.057</b>	<b>0.596</b>	<b>0.060</b>	<b>0.005</b>
	<b>US06</b>	<b>30</b>	<b>0.085</b>	<b>0.064</b>	<b>11.710</b>	<b>0.225</b>	<b>0.237</b>
	<b>US06</b>	<b>330</b>	<b>0.093</b>	<b>0.074</b>	<b>10.407</b>	<b>0.216</b>	<b>0.146</b>

Figure 1. Histogram of the NH<sub>3</sub> Emissions

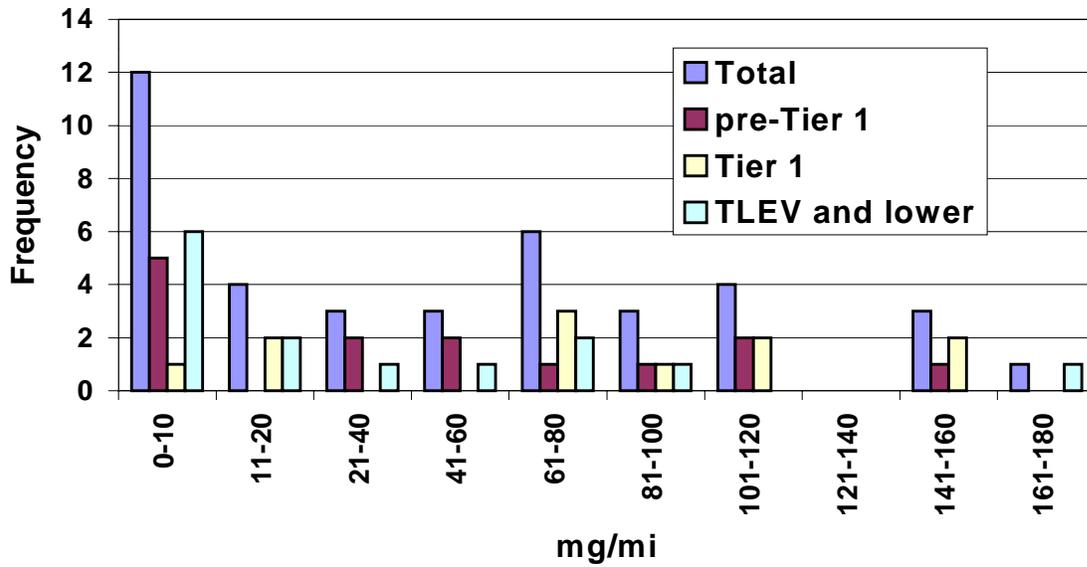


Figure 2. Average Emission Results for THC, CO, NO<sub>x</sub>, and NH<sub>3</sub>

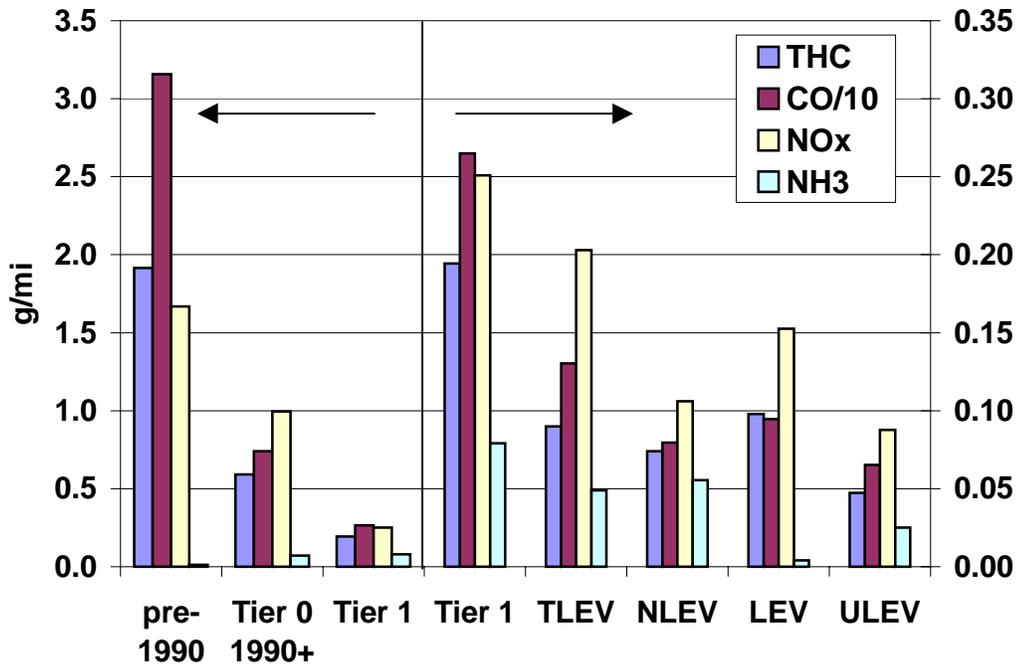


Figure 3. NH<sub>3</sub> Emissions for FTP, US06, NYCC and High Speed Freeway Cycles

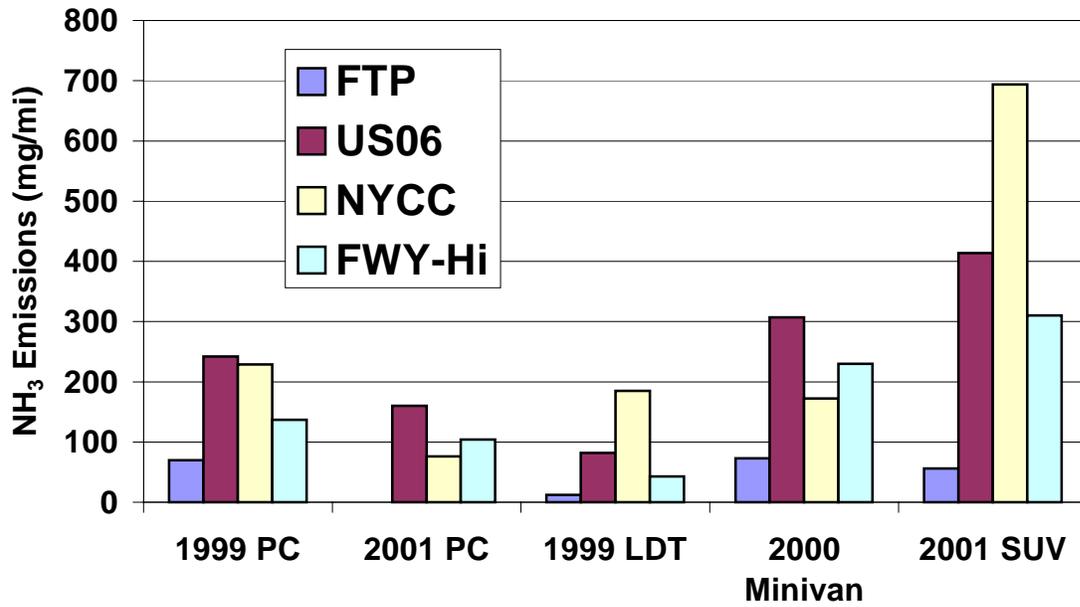


Figure 4. NH<sub>3</sub> vs Speed for FTP Cycle

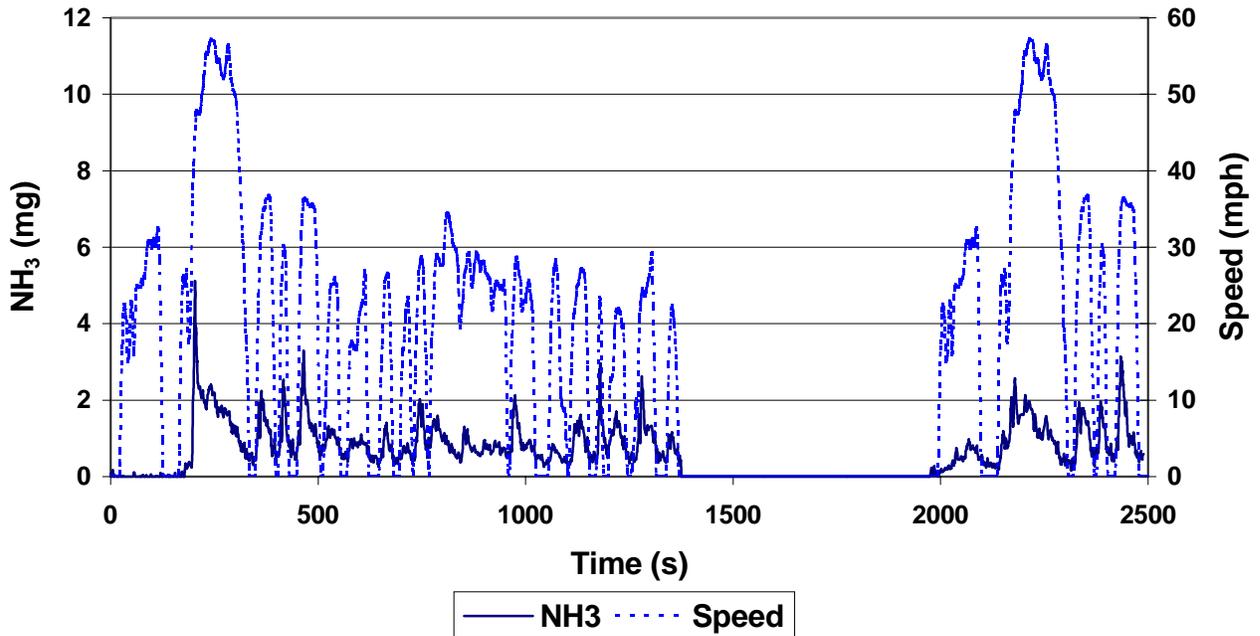
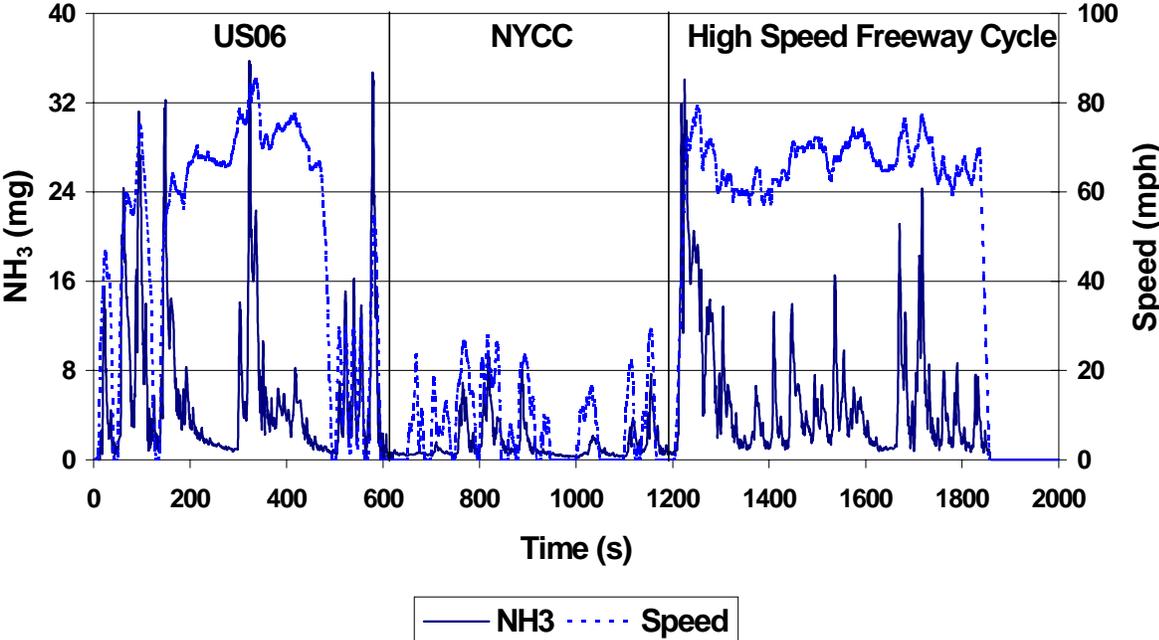


Figure 5 NH<sub>3</sub> vs Speed for US06, NYCC and High Speed Freeway Cycles



## Appendix A. FTP Results for Individual Vehicles

Number	vehicle	class	cert	odometer	Fuel	NH <sub>3</sub>	THC	NO <sub>x</sub>	CO
						g/mi	g/mi	g/mi	g/mi
1	1998 Honda Accord	PC	ULEV	2,402	CA RFG (in-use)	0.001	0.033	0.109	0.461
2	2000 Ford Windstar	LDT	ULEV	26,401	CA RFG (in-use)	0.073	0.036	0.077	0.489
3	2001 Buick LeSabre	PC	ULEV	9,986	CA RFG (in-use)	0.000	0.073	0.077	1.010
					ave	0.025	0.047	0.088	0.653
4	1997 Toyota Camry LE	PC	LEV	110,332	CA RFG (in-use)	0.005	0.122	0.146	0.943
5	1998 Nissan Altima GXE	PC	LEV	22,460	CA RFG (in-use)	0.003	0.074	0.159	0.949
					ave	0.004	0.098	0.153	0.946
6	2001 Suzuki Grand Vitara	SUV	NLEV	10,311	CA RFG (in-use)	0.056	0.074	0.106	0.795
7	1997 Ford Escort	PC	TLEV	5,403	CA RFG (cert)	0.038	0.054	0.058	0.514
8	1999 Toyota Tacoma	LDT	TLEV	30,204	CA RFG (in-use)	0.011	0.132	0.297	1.888
9	1999 GMC SonomaSL	LDT	TLEV	1,677	CA RFG (in-use)	0.000	0.044	0.079	0.426
10	1999 Dodge Stratus	PC	TLEV	19,080	CA RFG (in-use)	0.001	0.100	0.123	0.372
11	1998 Ford Ranger	LDT	TLEV	25,599	CA RFG (in-use)	0.088	0.103	0.233	1.471
12	1997 Honda Accord LX	PC	TLEV	75,132	CA RFG (in-use)	0.062	0.091	0.205	1.211
13	1999 GMC Sonoma	LDT	TLEV	10,610	CA RFG (in-use)	0.012	0.092	0.093	1.284
14	1996 Ford Explorer	SUV	TLEV	38,791	CA RFG (in-use)	0.177	0.103	0.535	3.257
					ave	0.049	0.090	0.203	1.303
15	1997 Dodge 3500 Van	LDT	Tier 1	40,678	CA RFG (in-use)	0.118	0.267	0.412	3.848
16	1996 Honda Civic	PC	Tier 1	58,755	CA RFG (in-use)	0.144	0.153	0.123	1.828
17	1996 Ford F150	LDT	Tier 1	133,524	CA RFG (in-use)	0.015	0.190	0.114	1.876
18	1999 Toyota Tacoma	LDT	Tier 1	27,289	CA RFG (in-use)	0.011	0.175	0.204	2.300
19	1995 Honda Civic	PC	Tier 1	108,008	CA RFG (in-use)	0.157	0.284	0.235	4.466
20	1998 Chevy S10	LDT	Tier 1	29,663	CA RFG (in-use)	0.070	0.194	0.287	5.353
21	1993 Ford ExplorerXLT	SUV	Tier 1	100,887	CA RFG (in-use)	0.095	0.203	0.410	2.959
22	1996 Honda Accord DX	PC	Tier 1	81,237	CA RFG (in-use)	0.076	0.106	0.258	1.479
23	1999 Ford Contour	PC	Tier 1	15,254	CA RFG (in-use)	0.070	0.090	0.152	0.807
24	1996 Ford Taurus	PC	Tier 1	68,154	CA RFG (in-use)	0.109	0.300	0.258	2.282
25	1997 GMC Sonoma	LDT	Tier 1	64,915	CA RFG (in-use)	0.007	0.176	0.306	1.951
					ave	0.079	0.194	0.251	2.650
26	1992 Dodge Spirit FFV	PC	1990+/Tier 0	15,264	CA RFG (cert)	0.118	0.177	0.176	2.791
replicate	1992 Dodge Spirit FFV		1990+/Tier 0		CA RFG (cert)	0.119	0.152	0.167	2.364
27	1994 Chevy Suburban	SUV	1990+/Tier 0	93,986	CA RFG (in-use)	0.046	1.660	2.041	17.029
28	1993 Nissan 240SX	PC	1990+/Tier 0	89,924	CA RFG (in-use)	0.084	0.286	0.617	4.555
29	1991 GMC Sonoma	LDT	1990+/Tier 0	103,245	CA RFG (in-use)	0.068	0.601	0.631	11.619
30	1993 Dodge Intrepid	PC	1990+/Tier 0	139,688	CA RFG (in-use)	0.036	0.528	1.030	4.049
replicate	1993 Dodge Intrepid		1990+/Tier 0		CA RFG (in-use)	0.018	0.609	1.023	4.475
31	1992 Toyota Corolla	PC	1990+/Tier 0	105,418	RFA	0.003	0.305	0.556	2.323
32	1990 Ford Thunderbird	PC	1990+/Tier 0	164,380	RFA	0.035	0.461	0.700	10.529
33	1991 Honda CRX	PC	1990+/Tier 0	166,205	CA RFG (in-use)	0.111	0.330	1.812	5.759
34	1993 Mazda Protégé LX	PC	1990+/Tier 0	106,782	CA RFG (in-use)	0.155	0.943	1.394	7.980
					ave	0.072	0.591	0.994	7.404

## Appendix A. FTP Results for Individual Vehicles

Number	Vehicle	class	cert	Odometer	Fuel	NH <sub>3</sub> g/mi	THC g/mi	NO <sub>x</sub> g/mi	CO g/mi
35	1989 Chevy G20	LDT	Pre-1990/Tier 0	88,771	RFA	0.064	1.155	0.748	12.021
replicate	1989 Chevy G20			88,771	RFA	0.052	1.355	0.850	11.328
36	1989 Chevy Celebrity	PC	Pre-1990/Tier 0	166,316	RFA	0.000	1.419	1.349	8.534
replicate	1989 Chevy Celebrity			166,316	RFA	0.000	1.492	1.373	9.089
37	1988 Chevy C-20 truck	LDT	Pre-1990/Tier 0	28,547	CA RFG (in-use)	0.000	0.533	0.734	5.241
38	1983 Olds Delta 88	PC	Pre-1990/Tier 0	95,541	CA RFG (in-use)	0.003	1.394	1.741	15.240
39	1968 Chevy C20	LDT	Non-catalyst	85,533	CA RFG (in-use)	0.000	4.938	3.709	116.980
					ave	0.012	1.915	1.669	31.589