



Fugitive Emissions from a Dry Coal Fly Ash Storage Pile

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Why Are Fugitive Particulate Emissions Important?



- Upcoming coal ash storage regulation may rely on dry ash handling
- Potential impact to local communities
- EPA proposing to lower annual $PM_{2.5}$ standard to $12 \mu\text{g m}^{-3}$.
- Detailed prevention of significant deterioration (PSD) analyses are required for a new/modified source when emissions for $PM_{2.5} > 10$ tons per year (tpy), for $PM_{10} > 15$ tpy and for TSP > 25 tpy.

Technical Motivation to Re-evaluate Fugitive Emissions

- Inaccurate fugitive emissions can lead to ineffective emissions control
- Fugitive EFs for fly ash highly uncertain
 - Ash handling, wind erosion or pile maintenance, transfer...
 - May not account for most important materials characteristics, site-specific data, or current materials handling practices



Overall Study Plan

- Phase 1: Dry fly ash– TVA Colbert Plant (AL) 1200MW
- Phase 2: Coal dust – TVA Gallatin Plant (TN) 1000MW
- Phase 3: TBD - Limestone/gypsum? Road dust? Fly ash in Western U.S.?



Project Goals

- Quantify fugitive particulate emissions at coal-fired power plants for different handling practices.
- Compare new emission rates with those from EPA AP-42 handbook.
- Utilities use to inform facility permitting.
- May help evaluate emission mitigation strategies

AP-42
FIFTH EDITION
JANUARY 1995

**COMPILATION
OF
AIR POLLUTANT
EMISSION FACTORS**

**VOLUME I:
STATIONARY POINT
AND AREA SOURCES**

Office Of Air Quality Planning And Standards
Office Of Air And Radiation
U. S. Environmental Protection Agency
Research Triangle Park, NC 27711

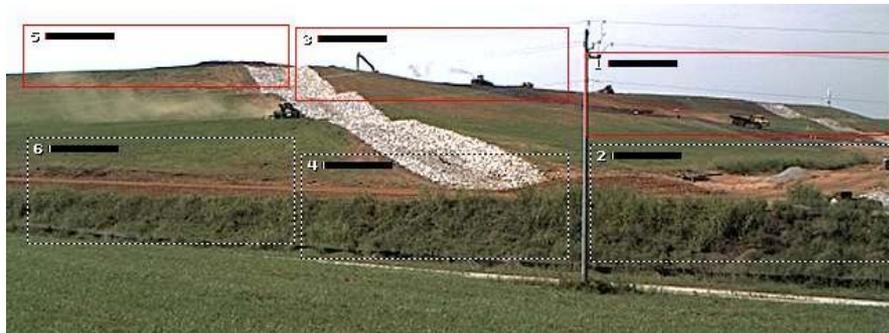
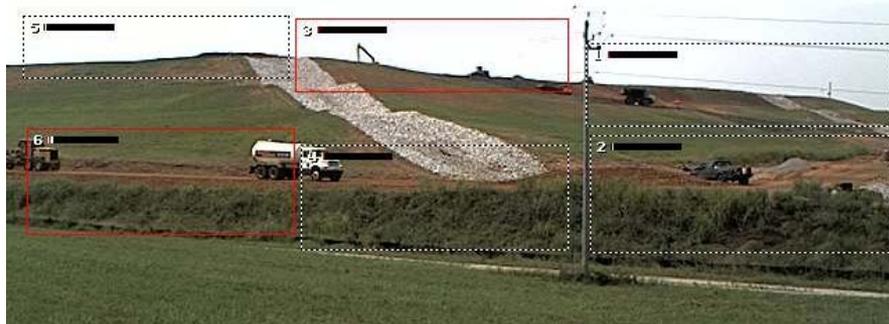
January 1995

Origins of AP-42 Fugitive Emission Factors

- Studies in 1970s measured airborne dust near unpaved roads and material handling. Multi-component statistical analyses to develop formulations. Dropping factors -1980s
- Used old measurement technologies.
- Most sources were staged & not done under actual operating conditions.
- Involved limited types of materials & limited range in conditions (vehicle speed, material moisture content, silt content).



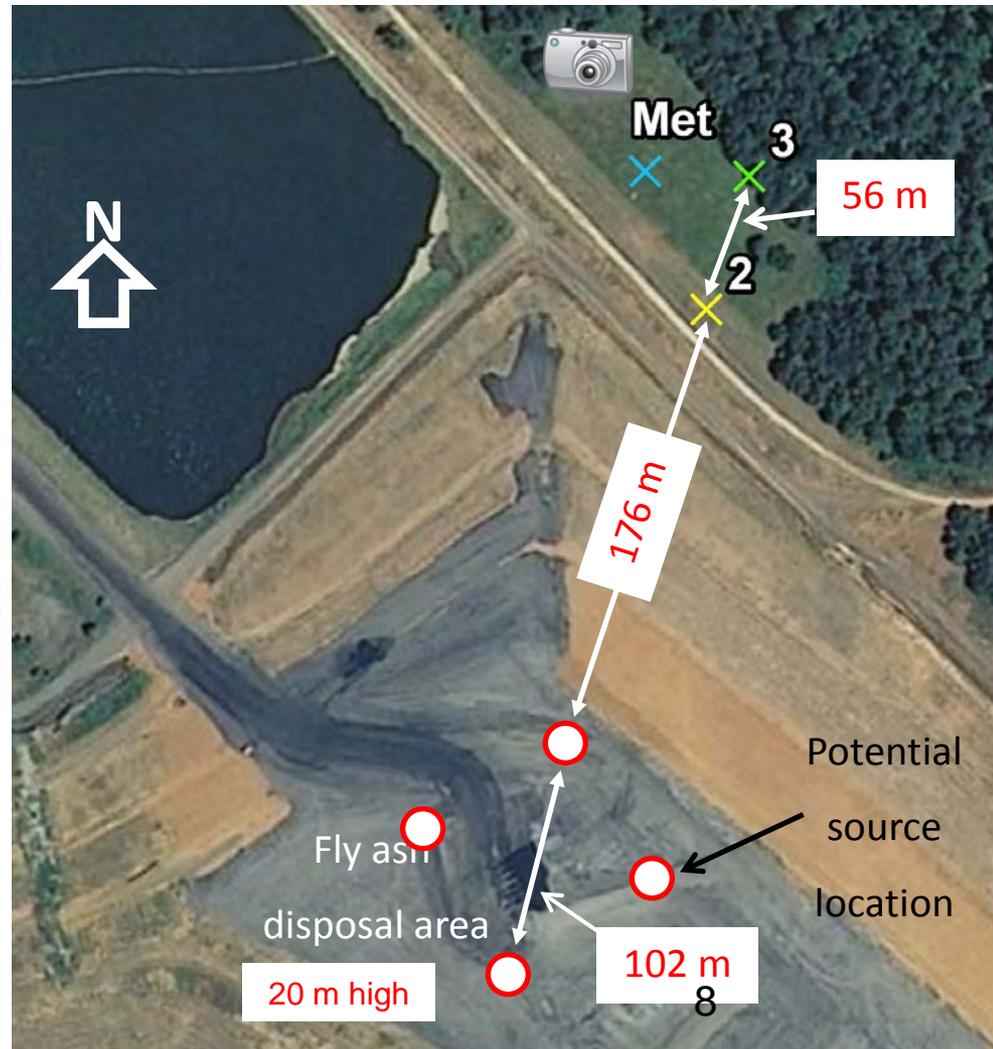
Colbert Field Study Layout



Photos:

(Top) Camera triggered by trucks moving along berm road south of air monitoring sites.

(Bottom) Camera triggered by activity on fly ash dry stack.



Monitoring Instrumentation

Meteorological

Instrument	Measurement/Purpose
R. M. Young 81000RE sonic anemometers (2 & 10 m)	Wind speed, direction, vertical velocity, horizontal & vertical turbulence; vertical gradient of speed, direction & turbulence
Vaisala HMI41 aspirated temperature & humidity sensors (2 & 10 m)	Air temperature & relative humidity; vertical temperature & moisture gradients
Campbell Scientific CNR2 net radiometer (~1.8 m)	Radiation flux (shortwave, longwave & net)
Novalynx Corp. 260-2501-A tipping bucket raingage	Precipitation amount
Campbell Scientific CS616 water content reflectometer (top 30 cm of soil)	Soil moisture content

Air Quality

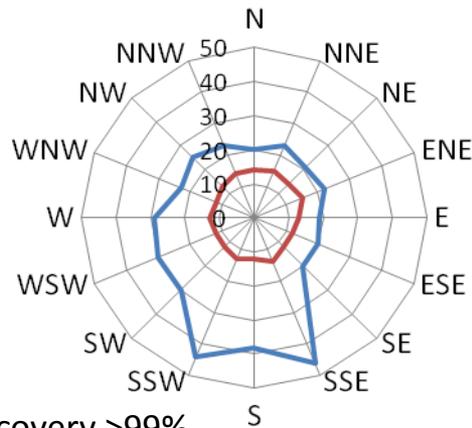
Instrument	Measurement/Purpose
Met One beta attenuation monitors (BAMs) at downwind & background sites	PM _{2.5} /PM ₁₀ concentrations (semi-continuous) @ downwind & background sites
BGI PQ-200 PM ₁₀ particle filter sampler	Filtered PM ₁₀ sample, ~12-hr samples (for chemical analysis) @ downwind & background sites
TSI 3563 3-λ nephelometer	Continuous β_{scat} @ 3 wavelengths @ downwind site
Optek nephelometer	Semi-continuous single-wavelength β_{scat} @ downwind site
Campbell Scientific video camera	Semi-continuous images of fly ash disposal site

Dry Fly Ash Handling at Colbert

- Fly ash is pneumatically conveyed to hoppers where it is conditioned with 15% moisture before transference to haul trucks.
- Each truck moves 25-28 m³ of ash per load.
- Ash is dropped at disposal area and leveled to a depth of about half a meter (18-24 in).
- Ash grading takes about 8 min with grader moving at 2.2 m s⁻¹ (5 mph).
- Fugitive emissions are primarily due to dropping and grading operations (haul trucks moving over bottom ash road produces very little fugitive dust).

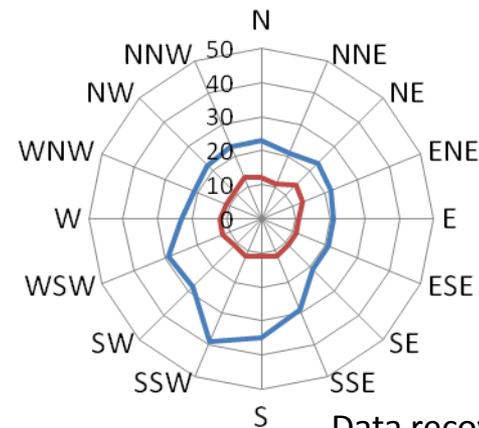
Average Particulate Concentrations by Direction May-August 2011

Nearest Downwind Site (2)



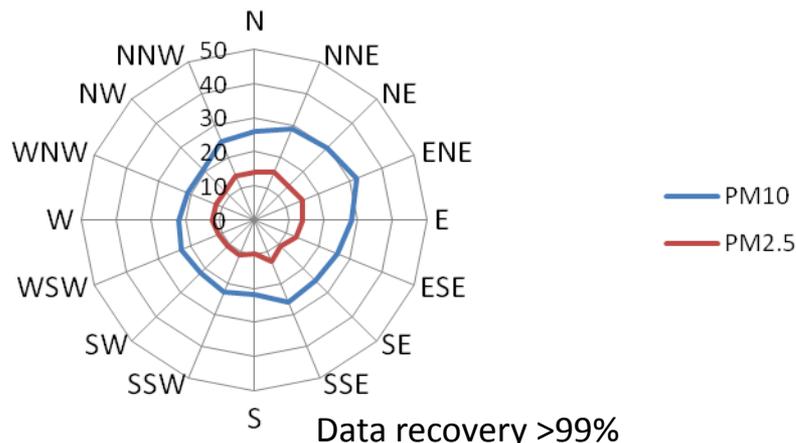
Data recovery >99%

Farthest Downwind Site (3)



Data recovery 81%

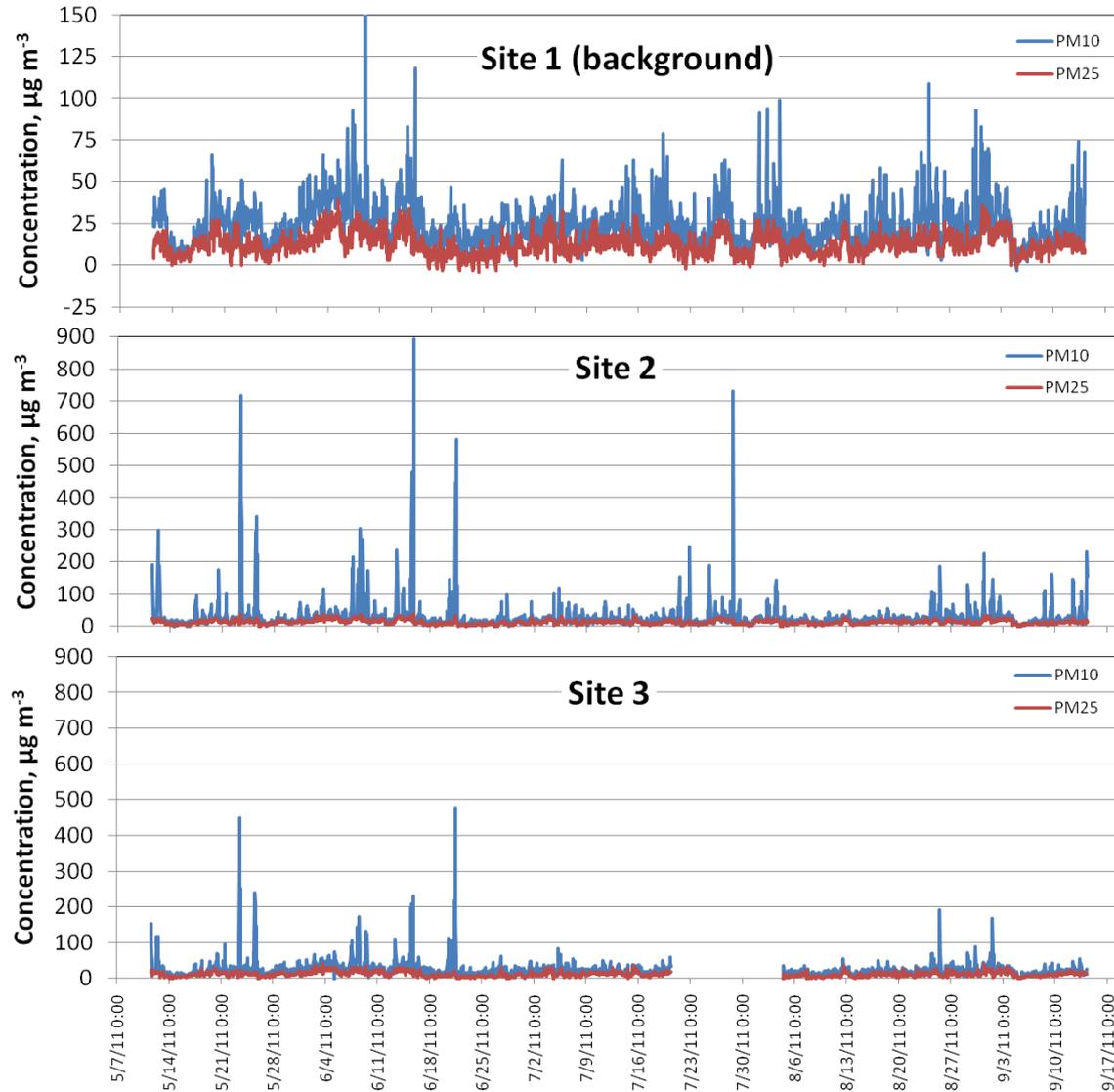
Background



Data recovery >99%

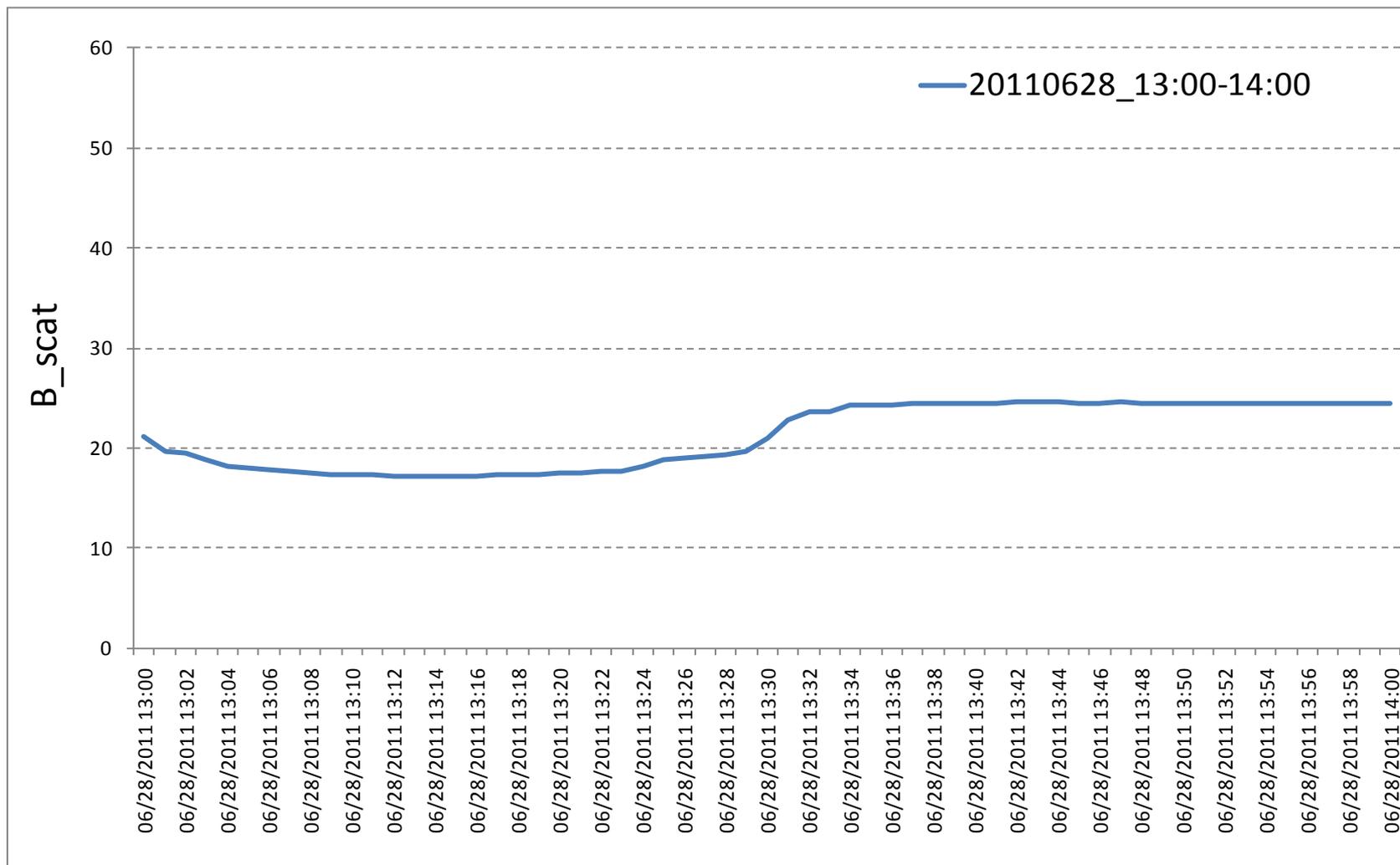
One-hr average concentrations were measured by FRM BAMs and reported here in $\mu\text{g m}^{-3}$. Wind direction was measured at a height of 10 m.

Hourly Concentrations at each Site: May-September 2011



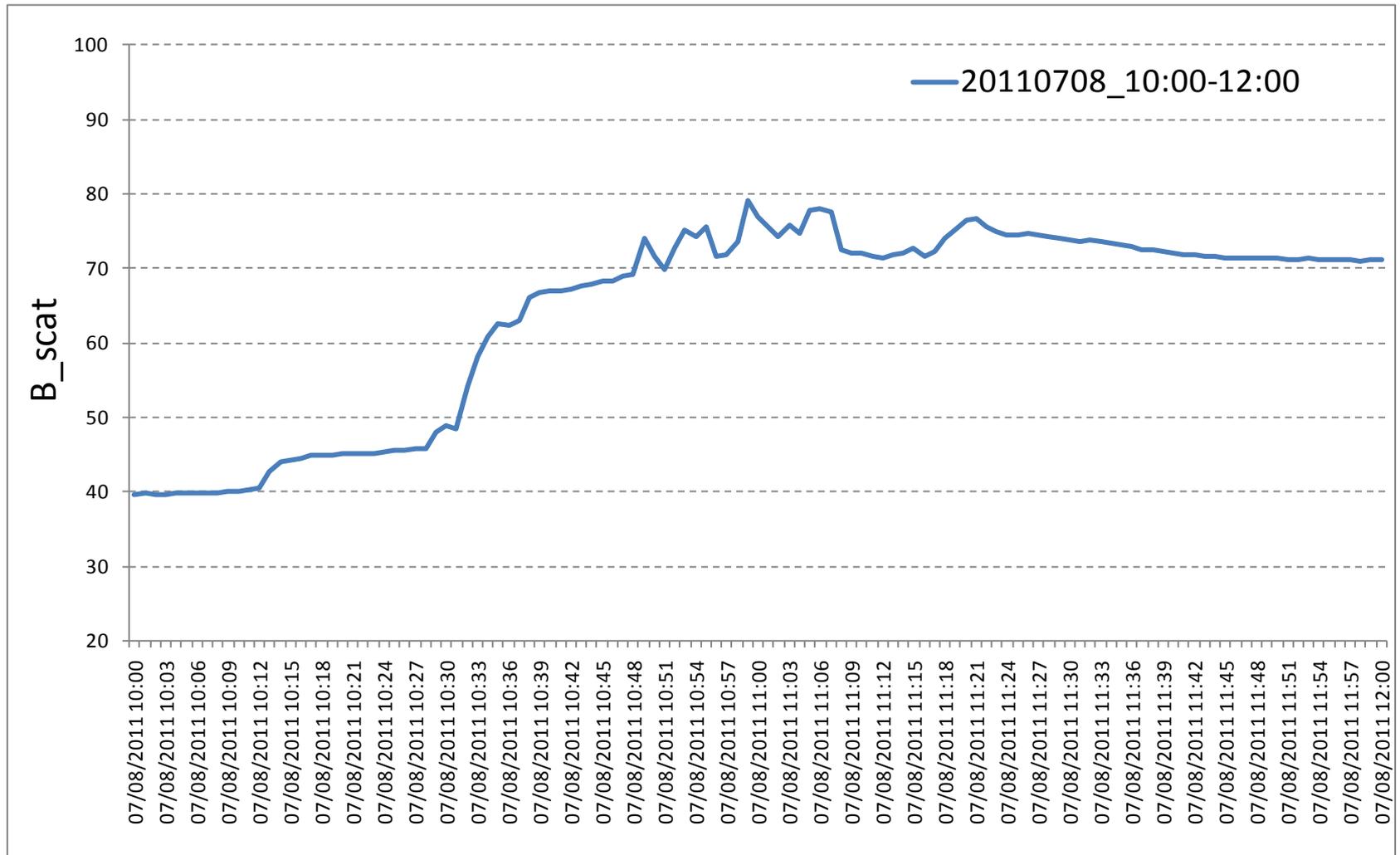
Clean Period (No Emissions)

28 June, 1300-1400 LST

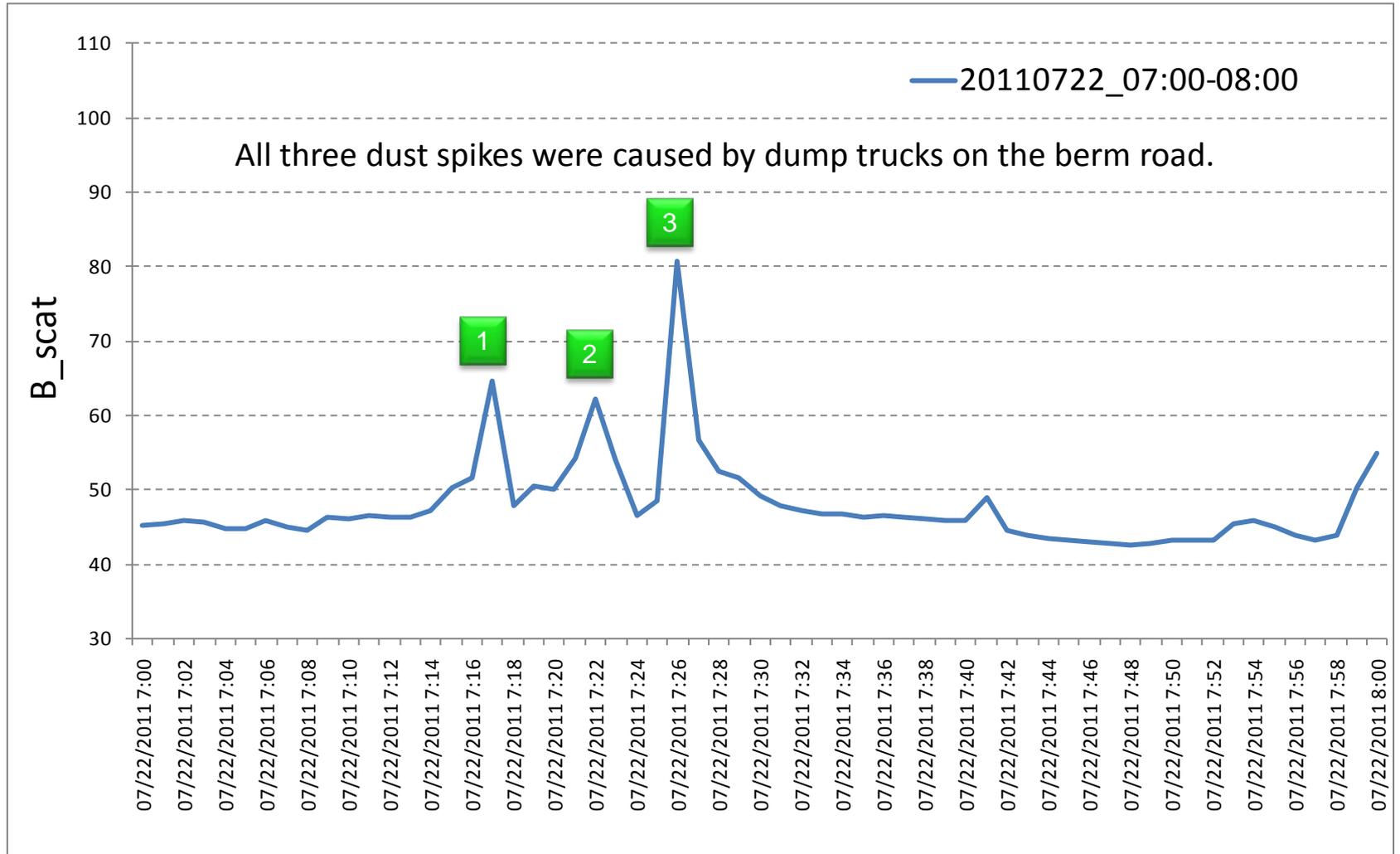


No-Vehicle Period (Fly Ash Activity only)

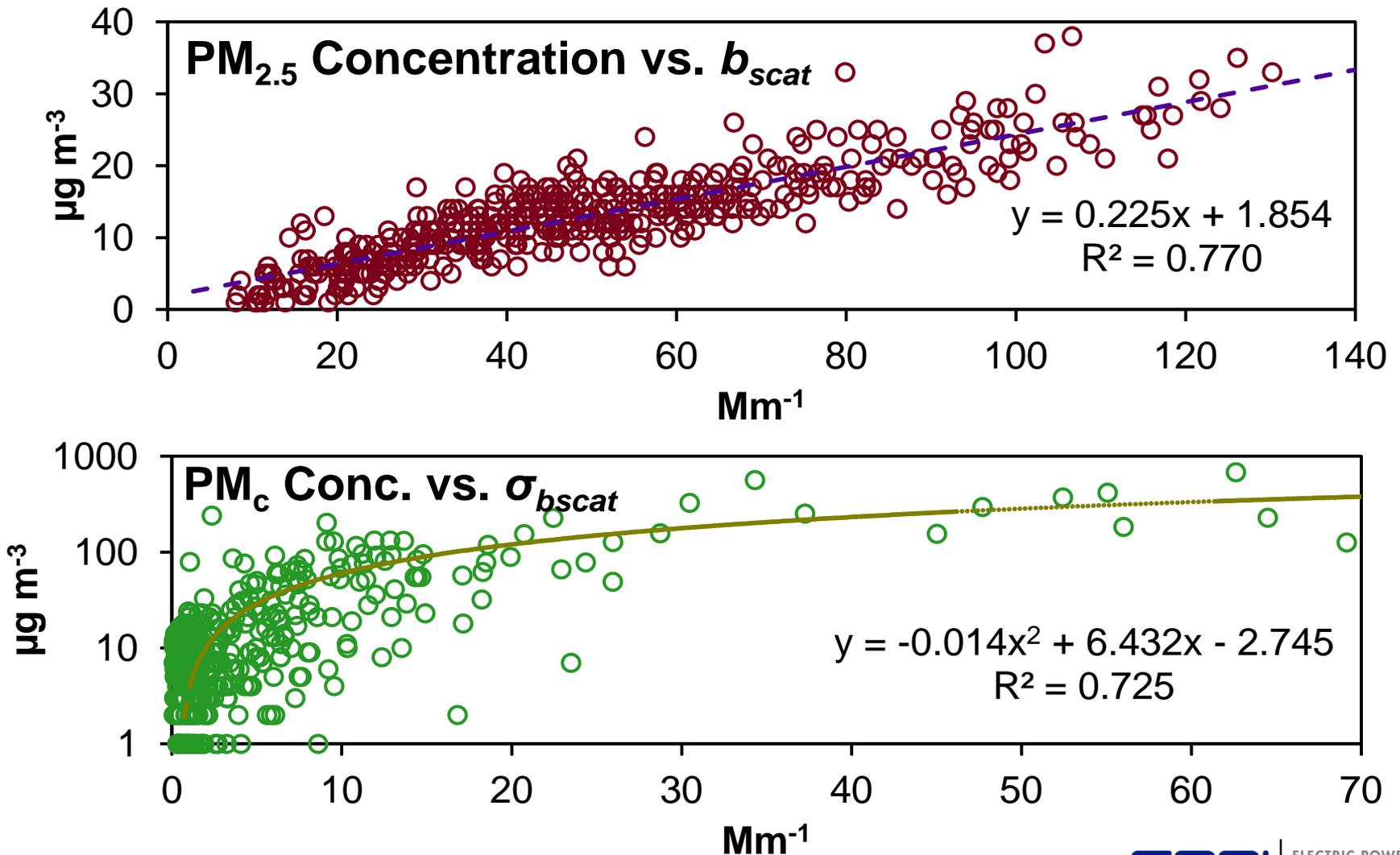
8 July, 1000-1200 LST



3-Vehicle Event 22 July, 0700-0800 LST



Particle Concentrations (SSE-SSW)



Models of Hourly PM_c & PM_{2.5}

Multivariate linear regression yields

$$C_{fine} = f_{b_{scat}} b_{scat} + f_{U_2} U_2 + C_{int} \quad r^2 = 0.89$$
$$C_{coarse} = f_{\sigma} \sigma_{b_{scat}} + f_{U_2} U_2 + f_{fine} C_{fine} + C_{int} \quad r^2 = 0.77$$

f : regression slope constants

b_{scat} : light scattering coefficient (Mm⁻¹)

σ_{scat} : standard deviation of b_{scat} (Mm⁻¹)

U_2 : wind speed at level 2 (m s⁻¹ @ 10 m)

C_{int} : intercept constants (μg m⁻³)

C_{fine} : concentration of PM_{2.5} mass (μg m⁻³)

C_{coarse} : concentration of PM_c mass (μg m⁻³)

Steps in Estimating Fly Ash Emission Rates

For each particle size:

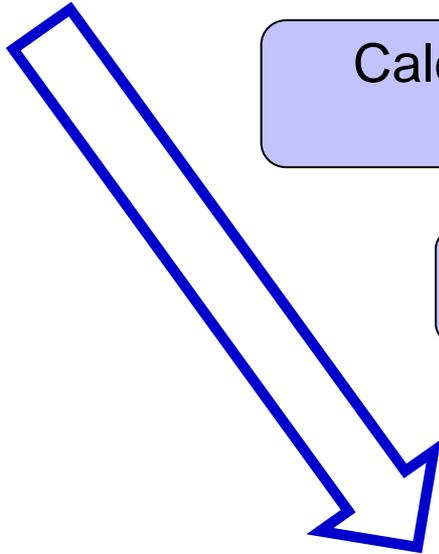
Select hours meeting minimum data requirements

Calculate “adjusted” concentrations at sites 2 & 3 to remove effects of nearby emissions

Calculate “excess” concentrations (i.e., values above background), C^{XS}

Compute normalized concentrations (C/Q)

Compute $Q_i = C^{XS} / (C/Q)$



Fly Ash Emission Equations

Based on AP-42:

$$E_{ash} = N_{loads} (E_{drop} + DE_{grad} / m)$$

N_{loads} = ash loads per unit time

D = grader distance traveled per load processed

m = ash mass per load

$$E_{drop} = C_{drop} \frac{(U/2.2)^{1.3}}{(M/2)^{1.4}}$$

$$E_{ir} = C_{ir} \left(\frac{S}{12}\right)^a \left(\frac{W}{3}\right)^{0.45} \quad \text{grading}$$

Based on field data & modeling:

$$C_{xsobs} = C_{obs} - C_{local} - C_{bck}$$

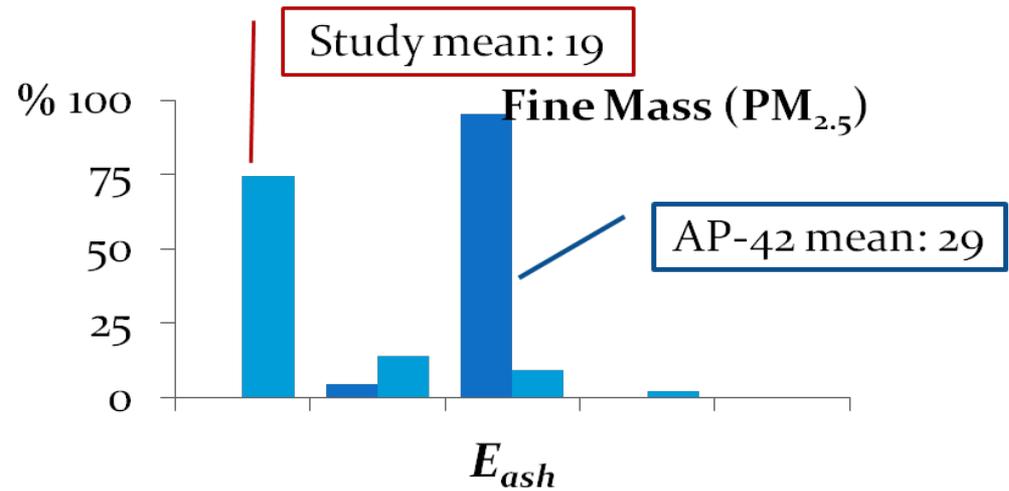
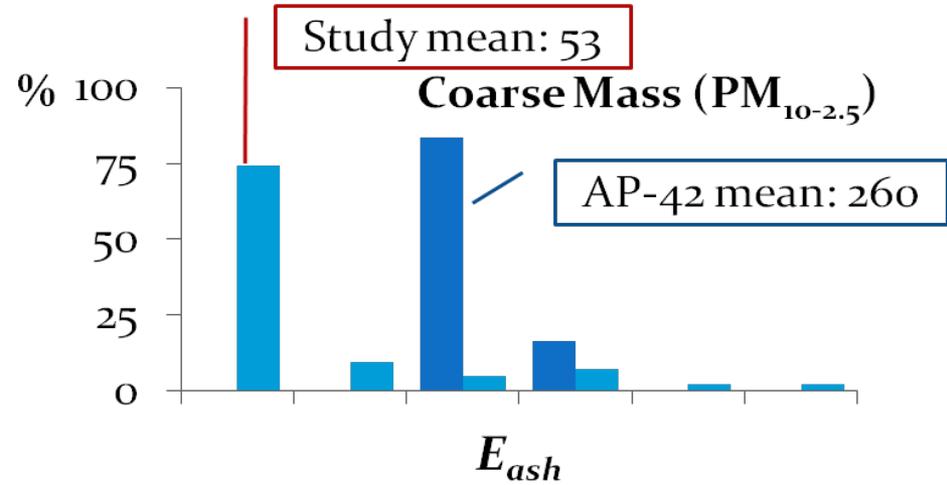
$$Q_{ash} = C_{xsobs} / (C/Q)_{model}$$

$$E_{ash} = Q_{ash} / M_{ash}$$

M_{ash} = ash mass deposited per area per time



Overall Emission Factors



g particles/ metric ton of processed ash

Estimating Vehicle Dust Emissions

Assuming an infinite line source and a Gaussian plume mass profile, calculate an emission rate of mass per unit length of road per unit time as (following Hanna et al., 1982)

$$Q_i = 2.46 C_i (Kux)^{1/2} \exp[uz^2/(4Kz)]$$

where C_i is observed concentration of mass component i , u is wind speed, x is downwind distance, K is eddy diffusivity, and z is the vertical height displacement from the plume centerline ($z=0$ for a ground level source). Near the ground under steady-state conditions, K can be determined using the relation

$$u_*^2 = K(du/dz)$$

with both u_* and du/dz known from measurements at the nearby meteorological tower.

Emission Factor Summaries

Emission Factor	Coarse Mass, PM _c		Fine Mass, PM _{2.5}	
	Mean	Median	Mean	Median
Fly Ash, AP-42 (g PM per Mg ash)	250	232	45	41
Fly Ash, this study (g PM per Mg ash)	63	12	14-18	5-7
Road dust, AP-42 - industrial sfc. (g PM per km traveled)	193	200	34	36
Road dust, AP-42 - public roads (g PM per km traveled)	32	26	5.7	4.6
Road dust, this study (g PM per km traveled)	68 ^e	38 ^e	3.3	1.4

What controls total ash disposal EFs?

- Grading activity accounts for >95 percent of total computed fugitive fly ash emissions.
- AP-42 grading EF formulation is for vehicles on *industrial roads* but was developed from surfaces relatively low (<25%) in silt content, no moisture, higher speeds.
- Tested both AP-42 unpaved road dust formulations at Colbert on 2 roads/42 events.
- Results*:
 - Industrial* AP-42 EF **10x** observed for PM_c
 - Industrial* AP-42 EF **23x** observed for $PM_{2.5}$
 - Public road* AP-42 EF **2x** observed for PM_c
 - Public road* AP-42 EF **4x** observed for $PM_{2.5}$

*Differences between AP-42 and field results are even larger when vehicle wake effects are considered.

Conclusions

- Despite conservative assumptions, AP-42 derived fly ash handling EFs are higher than EFs derived by field measurements for both PM_c and $PM_{2.5}$
- Disparity is likely due to high bias in industrial unpaved road dust formulation (grading).
- EFs from field measurements have higher tail than AP-42 EFs. May be due to higher variability in atmospheric or ash handling conditions in real operations.
- Use of more realistic EFs can lower fugitive dust emission estimates for fly ash handling by 33% ($PM_{2.5}$) to 80% (PM_{10}). *Can benefits be expanded to TSP?*
- Observed EFs for vehicles on unpaved roads also differ from AP-42 EFs.



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