



CASTNET



CASTNET Overview

The Clean Air Status and Trends Network (CASTNET) operated 100 monitoring stations throughout the contiguous United States, Alaska, and Canada in 2021. Ninety-four sites included filter pack systems to sample weekly concentrations of acidic gases and particulate matter. Eighty-eight sites included continuous ozone (O₃) analyzers. The U.S. Environmental Protection Agency (EPA), the National Park Service (NPS), and the Bureau of Land Management-Wyoming State Office (BLM) are the primary sponsors of CASTNET. NPS operated 31 sites during 2021. The BLM operated five sites in Wyoming. A map of CASTNET sites is shown in Figure 1.

Many organizations, including federal and state government agencies, Native American tribes, and universities, contribute to CASTNET (https://www.epa.gov/castnet/castnet-program-partners). These participants sponsor individual CASTNET sites and provide in-kind services that support the overall performance of the network. For example, these participants change weekly filter packs, operate and repair instruments, and perform general site maintenance. Many also provide the land for the CASTNET site.

Background

In 1990, the Clean Air Act Amendments (CAAA) added language that called for the EPA Administrator to establish a national monitoring network to assess relationships between decreasing emissions and trends in air quality, deposition, and ecological effects. Consequently, EPA established CASTNET. The fundamental mission of CASTNET is to measure changes in air quality that result from national and regional emission controls of criteria pollutants such as the Acid Rain Program (ARP).

CASTNET began with 50 sites using filter pack systems to sample weekly concentrations of acidic gases and particulate matter. Most sites included continuous O₃ analyzers. CASTNET has provided consistent measurements over three decades. Over time, enhancements have been added to increase the value of the program, including:

- 1987: The predecessor to CASTNET, the National Dry Deposition Network (NDDN), began collecting measurements at approximately 15 sites in the eastern United States.
- 1990: CASTNET filter pack measurements of sulfur and nitrogen pollutants were taken at 50 sites.
- 1994: NPS began its participation with the operation of 15 sites primarily in the western United States.
- 2002: The first tribal site was installed in 2002 in Oklahoma in conjunction with the Cherokee Nation. The seventh tribal site was installed in 2020 in partnership with the Confederated Tribes of the Umatilla Indian Reservation.

- 2007: The National Atmospheric Deposition Program (NADP) Ammonia Monitoring Network (AMoN) began. AMoN has grown to more than 100 sites with approximately 70 percent of AMoN sites situated at CASTNET locations.
- 2011: Ozone systems were upgraded at all EPA-sponsored sites to become regulatory compliant. As of 2021, there were 56 EPA-sponsored regulatory O₃ sites, 30 NPS sites, and 2 BLM sites.
- 2012: The first small-footprint sites, which do not require temperature-controlled shelters, were established in New York (2) and Vermont (1). Additional small-footprint sites were subsequently added for a total of 12 small-footprint sites in 2021.
- 2012: The BLM began its participation in CASTNET with five filter pack sites in Wyoming.
- 2021: As of December 31, 2021, 100 sites were operational; 64 of these sites are operated by the EPA, 31 by NPS, and 5 by the BLM Wyoming.

The ARP, which was established by the U.S. Congress under Title IV of the CAAA and started in 1995, was enacted to reduce emissions of sulfur dioxide (SO_2) and nitrogen oxides (NO_x) from fossil fuel fired electric generating units (EGUs). The ARP has produced significant reductions in emissions since 1995. Other air pollution control programs were instituted to further reduce emissions of SO_2 and NO_x . Examples of these programs and their implementation years include:

- 2000 Ozone Transport Commission NO_x Budget Program
- 2003 NO_x State Implementation Plan (SIP) Call NO_x Budget Trading Program
- 2009 Clean Air Interstate Rule
- 2015 Cross-State Air Pollution Rule (CSAPR)
- 2017 CSAPR Update

These emission control programs and their benefits are discussed on the EPA's power sector website (https://www.epa.gov/power-sector/progress-report-program-basics). As an example, implementation of the CSAPR Phase I began on January 1, 2015, and required 28 states in the eastern United States to significantly improve air quality by reducing EGU emissions that cross state lines and contribute to ground-level O_3 and fine particulate matter ($PM_{2.5}$) in downwind states. CSAPR was enacted under the Clean Air Act's "good neighbor" provision to address regional interstate transport of O_3 and $PM_{2.5}$ pollution to comply with the 1997 O_3 National Ambient Air Quality Standards (NAAQS) and the 2006 $PM_{2.5}$ NAAQS, which were the standards in effect at the time. The rule required reductions in annual SO_2 , annual NO_x , and O_3 season NO_x emissions.

The emission reductions achieved under these programs have resulted in a substantive improvement in air quality as demonstrated by air pollutant concentrations measured by CASTNET and other cooperating networks. By 2021, EGUs required to comply with the ARP and/or CSAPR had reduced their SO₂ emissions to 940,000 short tons and NO_x emissions to about 78,000 short tons, decreases of 92 and 87 percent, respectively, from 1990 levels.

Locations of Monitoring Sites

A map of CASTNET monitoring sites is provided in Figure 1. In 2021, 100 monitoring sites operated at 98 distinct locations. Native American tribes operated seven CASTNET sites in the central and northwestern United States. To estimate precision across the network, co-located sites operated at Mackville, KY (MCK131/231) and Rocky Mountain National Park, CO (ROM406/206). The ROM406/206 pair ensures consistency between EPA (ROM206) and NPS (ROM406). Of the two Rocky Mountain monitoring sites, ROM406 is specified as the regulatory monitoring site for O₃. NADP AMoN sites are also displayed in Figure 1. If the AMoN site is co-located with a CASTNET site, its site ID is displayed with the CASTNET ID. If it is not co-located, it is shown with a black dot. Site locations can also be viewed on the CASTNET website (https://www.epa.gov/castnet/castnet-site-locations).



Gothic (GTH161) - Gunnison, CO

Measurements Recorded at CASTNET Sites

During 2021, 94 CASTNET sites measured weekly ambient concentrations of acidic pollutants, base cations, and chloride (Cl⁻) using a 3-stage filter pack with a controlled flow rate (WSP, 2020; 2021; https://epa.gov/castnet/documents-reports). Gaseous pollutant concentrations include nitric acid (HNO₃) and sulfur dioxide (SO₂). Particulate concentrations include nitrate (NO₃⁻), ammonium (NH₄⁺), sulfate (SO₄⁻), magnesium (Mg²⁺), calcium (Ca²⁺), potassium (K⁺), sodium (Na⁺), and Cl⁻. The filter pack is replaced each Tuesday and shipped to the CASTNET analytical chemistry laboratory for analysis. Ambient temperature is measured at all sites in part to enable conversion of concentrations to local conditions.

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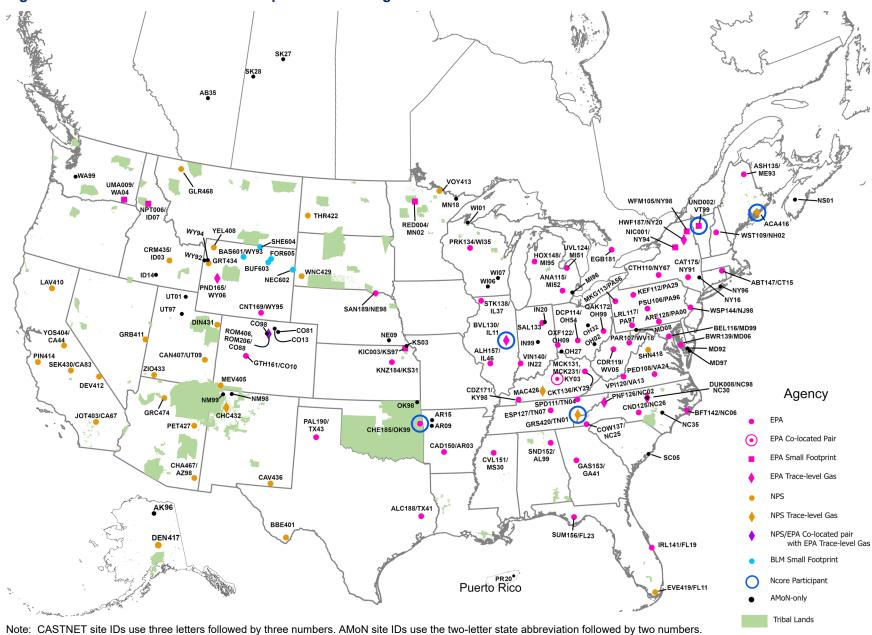


Figure 1 CASTNET and AMoN Sites Operational during 2021

4 CASTNET Overview

Most CASTNET sites include a temperature-controlled shelter and continuous O₃ monitoring system. CASTNET also deploys a small-footprint monitoring station that does not require a walk-in shelter. These sites can be operated using conventional or alternative power sources (wind/solar). The ability to use alternative power sources allows these sites to be placed in remote areas that are off the electrical grid. During 2021, CASTNET operated 12 small-footprint sites. Ozone concentrations were measured at 88 of the 100 sites. The O₃ inlet and filter pack are located atop a 10-meter (m) tower. Some CASTNET sites also measure trace-level NO/total nitrogen oxides (NO_x), SO₂, carbon monoxide, and nitric oxide/total reactive oxides of nitrogen (NO/NO_y). In 2021, meteorological parameters were measured at 5 EPA-, 31 NPS-, and 5 BLM-sponsored CASTNET sites. Measured meteorological parameters include 2-m temperature, wind speed and direction, standard deviation of the wind direction, solar radiation, relative humidity, precipitation, and surface wetness (at select sites).

Trends in fourth highest daily maximum 8-hour average (DM8A) O₃ concentrations and annual sulfur and nitrogen levels are depicted in the following sections using box plots showing 1-year mean, median, and annual distributions of concentrations aggregated from 34 eastern reference sites for 1990 through 2021 and from 16 western reference sites for 1996 through 2021. The reference sites were selected for their long-term data records and consistent performance. All data used to create the figures and maps in this report are available on the CASTNET website: https://www.epa.gov/castnet/download-data.

Related Air Quality Networks

CASTNET monitors air quality and deposition in cooperation with other long-term national and international networks, which produce data that EPA uses to assess the effectiveness of emission control programs and to characterize rural air quality nationwide. Although not shown in Figure 1, CASTNET sites are often co-located with one or more of the networks listed below. This includes the NADP (https://nadp.slh.wisc.edu/) and its other networks operated in addition to AMoN:

- National Trends Network (NTN)
- Mercury Deposition Network (MDN)
- Atmospheric Mercury Network (AMNet)

Other cooperating networks include:

- Canadian Air and Precipitation Monitoring Network (CAPMoN)
 (https://www.canada.ca/en/environment-climate-change/services/air-pollution/monitoring-networks-data/canadian-air-precipitation.html)
- EPA's National Core Monitoring (NCore) (https://www.epa.gov/amtic/ncore-monitoring-network)
- BLM's Wyoming Air Resources Monitoring System (WARMS) (https://www.blmwarms.net/)
- Interagency Monitoring of Protected Visual Environments (IMPROVE) (http://vista.cira.colostate.edu/IMPROVE/)

Building Tribal Partnerships with CASTNET Monitoring Sites

EPA partners with Cherokee Nation in Oklahoma (CHE185), the Alabama-Coushatta Tribe in Texas (ALC188), the Santee Sioux Tribe in Nebraska (SAN189), Kickapoo Tribe in Kansas (KIC003), the Red Lake Band of Chippewa Indians in Minnesota (RED004), the Nez Perce Tribe in Idaho (NPT006), and the Confederated Tribes of the Umatilla Indian Reservation in Washington (UMA009) to operate CASTNET sites on tribal lands. Most tribal sites utilize a small-footprint monitoring station that does not require a temperature-controlled shelter. The small-footprint design has produced cost savings and increased efficiency in terms of installation and power. EPA is interested in supporting additional tribal groups that want to establish small-footprint sites to monitor O₃ concentrations and assess sulfur and nitrogen inputs to their surrounding ecosystems.



Palo Duro (PAL190) - Randall, TX

Air Quality Impacts Due to Climate Change CASTNET provides a consistent data record that can be used to evaluate how short-term (extreme precipitation) and long-term (increasing temperature) changes impact air quality and the environment. CASTNET data are used to assess how and where emission reduction programs may become less effective due to climate change. CASTNET provides air quality data to communities that often lack state, local, or tribal monitors but have been disproportionately impacted by environmental or socioeconomic burdens. EPA has committed to reducing pollution within environmental justice communities. Figure 2 shows CASTNET sites in red that are located within or near (within 20 km) disadvantaged communities. These were identified using the Council on Environmental Quality's Climate and Economic Justice Screening Tool (https://www.epa.gov/ejscreen).

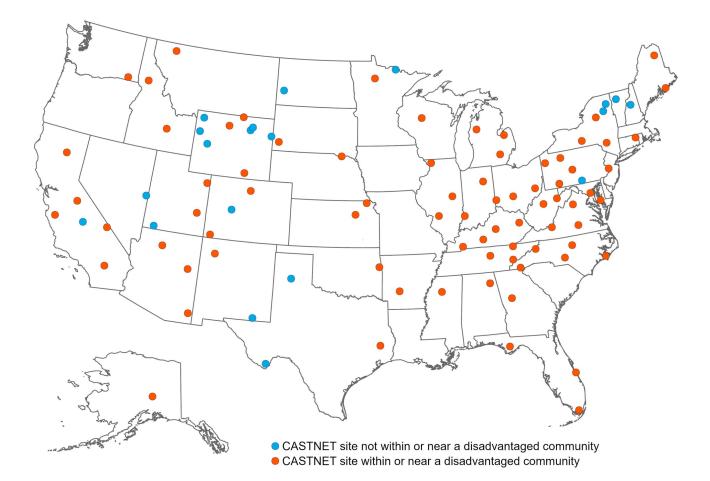


Figure 2 CASTNET Sites Within or Near Disadvantaged Communities

Quality Assurance Program

The CASTNET quality assurance (QA) program was established to ensure that all reported data are of known and documented quality in order to meet CASTNET objectives. The QA program also ensures intra-network consistency and comparability and the delivery of data that are reproducible and comparable with data from other monitoring networks. The 2021 QA program elements are documented in the CASTNET Quality Assurance Project Plan (QAPP; WSP, 2020; 2021). The QAPP includes standards and policies for all components of project operation, from site selection through final data reporting, with appendices that provide standard operating procedures for CASTNET operations. QA program metrics are available in quarterly and annual QA reports posted on the EPA CASTNET website: https://epa.gov/castnet/documents-reports.

Ozone Concentrations

CASTNET O₃ measurements are subject to rigorous QA/quality control procedures specified by federal regulations and the CASTNET QA Program. Hourly average concentrations were measured at 88 CASTNET sites in 2021. Data from 87 sites were delivered routinely to the EPA Air Quality System (AQS) database (https://www.epa.gov/aqs). One site, Duke Forest, NC (DUK008), is a research site operated by EPA's Office of Research and Development and does not meet regulatory siting criteria. The O₃ inlet at DUK008 is located above the forest canopy near the top of a 30 m tower. Two co-located monitors were designated as NAAQS-excluded, and the associated data from these monitors were used for QA purposes only. With the exception of DUK008 and the two NAAQS-excluded sites, data were used to calculate the three-year average of the fourth highest DM8A O₃ concentrations if three years of Title 40 Code of Federal Regulations Part 58-compliant data were available (Figure 3). CASTNET measurements provide data to support comparison with the O₃ NAAQS and to assess long-term trends and geographic patterns in regional O₃.

The primary O₃ NAAQS is designed to protect public health. The secondary standard is designed to protect public welfare and the environment. On October 1, 2015, the EPA promulgated a revision to the O₃ NAAQS, lowering the level of both the primary and secondary standards to 0.070 parts per million (ppm; https://www.federalregister.gov/documents/2015/10/26/2015-26594/national-ambient-air-quality-standards-for-ozone). On December 23, 2020, EPA decided to retain the 2015 O₃ NAAQS without changes to the form or level. To attain the 2015 standard, the three-year average of the fourth highest DM8A O₃ concentrations measured at each monitor within a specified area must not exceed 0.070 ppm or 70 parts per billion (ppb) in practice. For this report, the 2015 standard and related calculations were used. The design values for each standard can be downloaded from the EPA Air Quality Design Values website (https://www.epa.gov/air-trends/air-quality-design-values).

The EPA and other federal, tribal, state, and local agencies measure hourly O_3 concentrations through national and local monitoring programs. Measurements potentially affected by exceptional events, i.e., unusual or natural events that can affect air quality but are not reasonably controllable, were not removed when calculating the fourth highest daily maximum values for this report (https://www.epa.gov/air-quality-analysis/exceptional-events-rule-and-guidance).

Three-year averages of the fourth highest DM8A O_3 concentrations for 2019 through 2021 are presented in Figure 3. Ozone concentrations were not included on the map if the three-year average was not available because of incomplete data; these sites are shown as dots with no value. Table 1 lists sites with 2019–2021 O_3 design values exceeding 70 ppb, the level of the 2015 O_3 NAAQS.

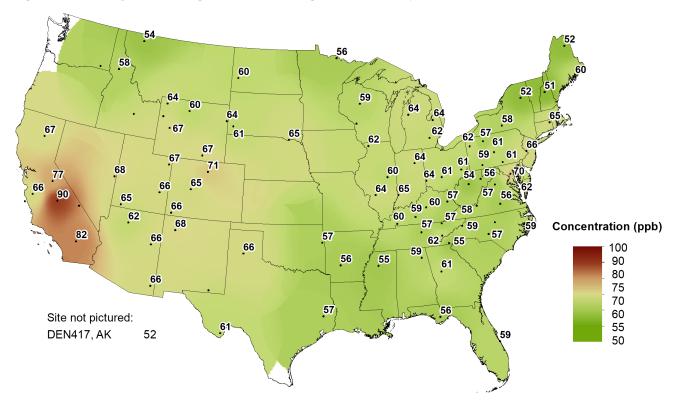


Figure 3 Three-year Averages of Fourth Highest DM8A O₃ Concentrations for 2019–2021

Table 1 2019–2021 O₃ Design Values Greater than 70 ppb

Site ID	State	Sponsor	3-year Average
SEK430	California	NPS	90
JOT403	California	NPS	82
YOS404	California	NPS	77
ROM406	Colorado	NPS	71

Figure 4 shows 2021 fourth highest DM8A O₃ concentrations that met percent completeness criteria.

Figure 5 provides box plots depicting trends in 1-year mean, median, and annual distributions of fourth highest DM8A O_3 concentrations from the 34 CASTNET eastern reference sites (Figure 5, 1 of 2) and the 16 western reference sites (Figure 5, 2 of 2).

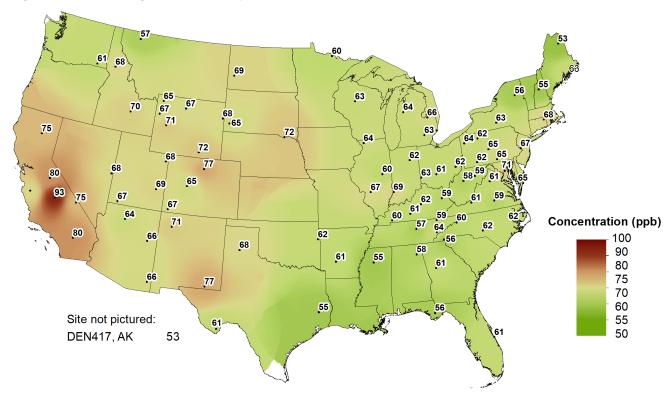
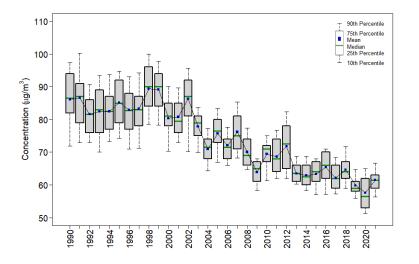


Figure 4 Fourth Highest DM8A O₃ Concentrations for 2021

Figure 5 Trends in Fourth Highest DM8A O₃ Concentrations (1 of 2)



Eastern Reference Sites

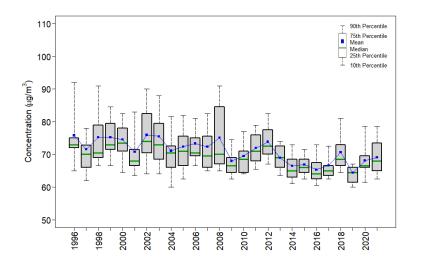


Figure 5 Trends in Fourth Highest DM8A O₃ Concentrations (2 of 2)

Western Reference Sites

Temperature Trends

Extensive scientific studies have demonstrated long-term increases in temperature across the United States and the world. Temperatures are rising, and rainfall and snow patterns are shifting. Extreme weather events such as heavy rainstorms and record high temperatures are becoming more common. Many of these observed weather and climate changes are linked to rising levels of greenhouse gases in the atmosphere. EPA provides information on temperature, precipitation, and other indicators that provide evidence of these long-term changes and their impacts on people and the environment (https://www.epa.gov/climate-indicators). EPA partners with dozens of data contributors (e.g., National Oceanic and Atmospheric Administration [NOAA]) to compile and keep these indicators up to date.

CASTNET sites were selected to characterize air quality in rural, regionally representative areas to avoid the undue influence of significant sources of air pollution from power plants and urban areas. The CASTNET eastern reference sites have been operating since 1990, and the CASTNET western reference sites have been operating since 1995. Therefore, CASTNET temperature measurements from 1990 through 2021 for the eastern reference sites and 1995 through 2021 for the western reference sites were selected to illustrate rural temperature changes. Seasons were defined as winter (December, January, February); spring (March, April, May); summer (June, July, August); and fall (September, October, November).

Temperature data were collected hourly at each site over the periods 1990 through 2021 (eastern sites) or 1995 through 2021 (western sites). The sites were divided into six geographic regions (Figure 6). The eastern sites were sorted into the Midwest, Southeast, and Northeast regions. The western sites were sorted into the Pacific, Rocky Mountain, and Southwest regions. Table 2 lists the

sites by region with the data collection period for each site. Each site produced more than 200,000 hourly temperature values.

To ensure these data sets were as complete and rigorous as possible, the data were analyzed for missing data and outliers. Missing data might affect calculation of trends and other analyses and, if extensive, might influence conclusions. Data were estimated for gaps in months that had less than 70 percent of the expected number of valid hourly temperature values for a given month with leap years taken into consideration. Linear interpolation was used on data gaps by the hour for a given month, day, and site location. The analysis was then done on the annual averages once the hourly data gaps were replaced. Outliers were caused by occasional system and instrument calibrations and power failures. They were detected and removed. The resulting temperature databases were considered sufficient for trends analyses.

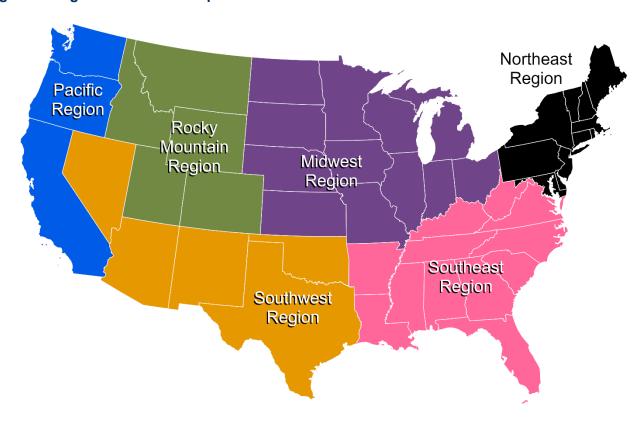


Figure 6 Regions Used for Temperature Trends

Note: Site MEV405, CO was included in the Southwest region.

Table 2 CASTNET Sites Used in Temperature Trend Analyses by Region

Cito ID	Range of	Percent				
Site ID Years Missing Data Pacific						
JOT403, CA	1995–2021	0.91				
LAV410, CA	1995–2021	2.41				
PIN414, CA	1995–2021	1.22				
YOS404, CA	1995–2021	5.73				
Rocky Mountain						
CAN407, UT	1995–2021	2.29				
CNT169, WY	1995–2021	3.89				
GLR468, MT	1995–2021	3.92				
GTH161, CO	1995–2021	1.94				
PND165, WY	1995–2021	3.89				
ROM406, CO	1995–2021	1.77				
YEL408, WY	1995–2021	6.67				
	Southwest					
BBE401, TX	1995–2021	3.35				
CHA467, AZ	1995–2021	1.21				
GRC474, AZ	1995–2021	4.00				
GRB411, NV	1995–2021	7.76				
MEV405, CO	1995–2021	1.68				
Midwest						
ALH157, IL	1990–2021	7.47				
ANA115, MI	1990–2021	3.28				
BVL130, IL	1990–2021	8.07				
DCP114, OH	1990–2021	6.03				
OXF122, OH	1990–2021	1.41				
PRK134, WI	1990–2021	1.97				
SAL133, IN	1990–2021	2.84				
UVL124, MI	1990–2021	2.88				
VIN140, IN	1990–2021 1.76					

Site ID	Range of Years	Percent Missing Data			
Cito ID	Southeast	Wilsoning Bata			
CAD150, AR	1990–2021	4.09			
CDR119, WV	1990–2021	3.37			
CND125, NC	1990–2021	6.43			
COW137, NC	1990–2021	0.85			
ESP127, TN	1990–2021	4.51			
GAS153, GA	1990–2021	3.58			
MCK131, KY	1990–2021	6.29			
PAR107, WV	1990–2021	0.66			
PED108, VA	1990–2021	3.00			
PNF126, NC	1990–2021	6.59			
SHN418, VA	1990–2021	2.36			
SND152, AL	1990–2021	3.12			
SPD111, TN	1990–2021	5.10			
SUM156, FL	1990–2021	2.66			
VPI120, VA	1990–2021	3.32			
Northeast					
ARE128, PA	1990–2021	0.79			
ASH135, ME	1990–2021	2.34			
BEL116, MD	1990–2021	4.45			
CTH110, NY	1990–2021	2.43			
KEF112, PA	1990–2021	3.16			
LRL117, PA	1990–2021	3.22			
MKG113, PA	1990–2021	3.30			
PSU106, PA	1990–2021	0.23			
WSP144, NJ	1990–2021	3.81			
WST109, NH	1990–2021	2.58			



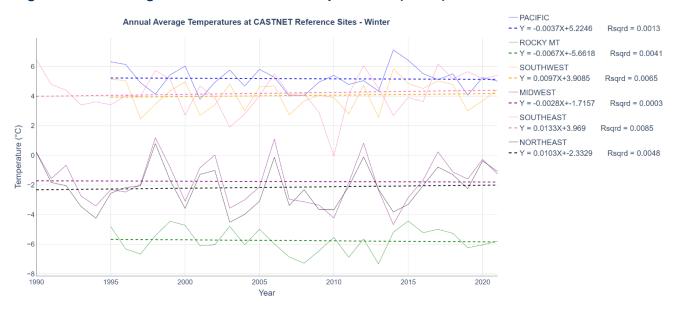
Speedwell (SPD111) - Claiborne, TN

Linear regression was used to analyze regional temperature trends by season. The linear regression preliminary results for the four seasons (winter, spring, summer, and fall) are presented in Figure 7. Winter regression analyses show decreasing trends for three regions (Pacific, Rocky Mountain, and Midwest). The figure shows increasing temperatures for all regions for spring, summer, and fall. Further examination would be needed to determine how best to utilize these data in the context of the longer-term temperature data.

Figure 8 illustrates total temperature change by season. Temperature change was calculated as the estimated slope times the range of years. The range of years for Pacific, Rocky Mountain, and Southwest was 27. For Midwest, Southeast, and Northeast, it was 32.

The annual average was calculated by grouping the temperature values of all the sites in each specified region. The assumption made was that all the sites are evenly distributed spatially. Future work will have to be done to calculate a more accurate annual average based on sites that are not evenly spaced. In addition, it is recommended that seasonal trend analysis is done on a data set that covers more than 30 years. This is especially true when looking at small geographical areas in a short time frame. Analyzing temperature trends over shorter time periods is difficult and is not consistent when compared with well-known, longer-term temperature trends. This report provides an overview of the temperature trends solely for data CASTNET has currently collected. CASTNET reports will continue to provide data and updated analyses in the future.

Figure 7 Linear Regression of Seasonal Temperatures (1 of 2)



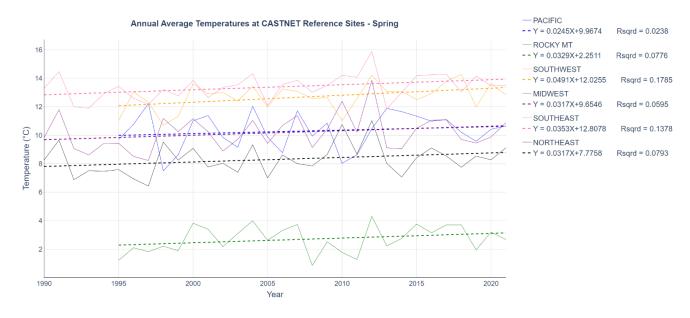
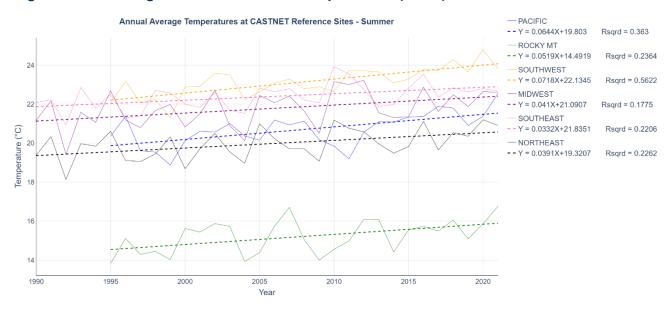
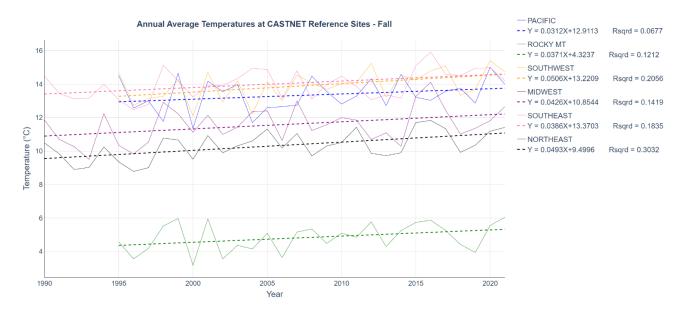


Figure 7 Linear Regression of Seasonal Temperatures (2 of 2)





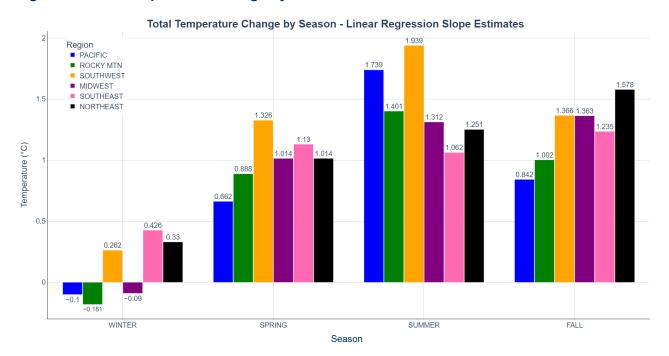


Figure 8 Total Temperature Change by Season

Nitrogen Atmospheric Concentrations

Annual mean concentrations of total NO₃ (HNO₃ + NO₃), NH₄, and NH₃ for 2021 are presented in three maps. Trends in annual mean concentrations of total NO₃ and NH₄ over the 32-year period, 1990 through 2021, were calculated from the eastern reference sites and for the 26-year period, 1996 through 2021, from the western reference sites. Trends in NH₃ were estimated from AMoN data from 2008 through 2021.

Trace-level gas analyzers measured NO_y at five EPA-sponsored and two NPS-sponsored CASTNET sites during 2021. These are presented at the end of this section. Measurements were not made at DUK008, NC during January through November 2021 due to the collapse of the Duke Forest hardwood tower in December 2020. Filter pack sampling restarted at DUK008 in December 2021.

Total Nitrate Concentrations

Nitric acid and particulate NO₃ are measured on CASTNET filter packs, and the sum is reported as total NO₃. Annual mean total NO₃ concentrations measured in 2021 are presented in the map in Figure 9. Sites that did not meet the completeness criterion appear as a dot with no value on the various concentration maps. To illustrate trends, Figure 10 provides box plots of total NO₃ levels for the eastern and western reference sites through 2021.

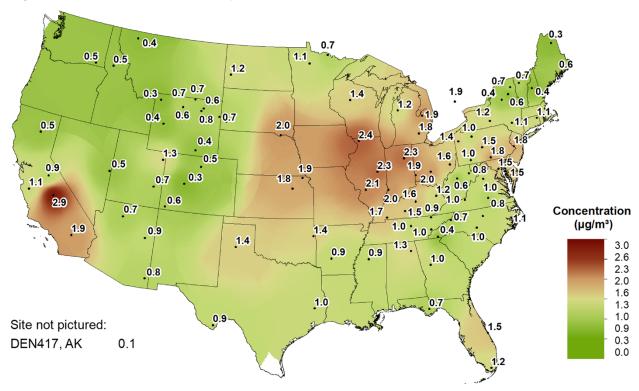
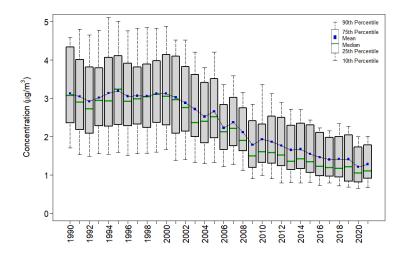


Figure 9 Annual Mean Total NO₃ Concentrations for 2021

Figure 10 Trends in Annual Mean Total NO₃ Concentrations (1 of 2)



Eastern Reference Sites

2.5 - 90th Percentile 7 Sth Percentile Mean Median Median Median 10th Percentile 10th Percenti

Figure 10 Trends in Annual Mean Total NO₃ Concentrations (2 of 2)

Western Reference Sites

Particulate Ammonium Concentrations

A map of 2021 annual mean particulate NH_4^+ concentrations is provided in Figure 11. Figure 12 shows box plots of NH_4^+ concentrations.

Figure 11 Annual Mean NH₄ Concentrations for 2021

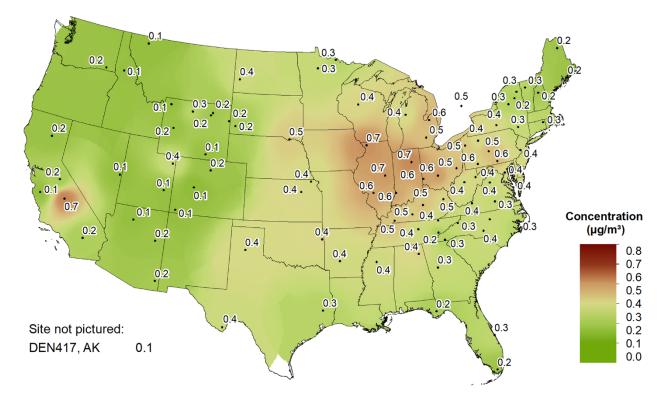
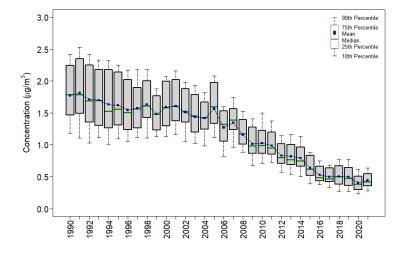
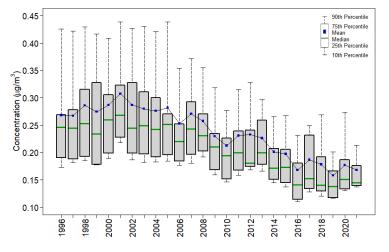


Figure 12 Trends in Annual Mean NH₄ Concentrations



Eastern Reference Sites



Western Reference Sites

Ammonia Concentrations

Annual mean NH₃ concentrations for 2021 from NADP/AMoN sites that met completeness requirements are mapped in Figure 13. Trends from NADP/AMoN sites that were active from 2008 through 2021 are shown in Figure 14.



Sumatra (SUM156) - Liberty, FL

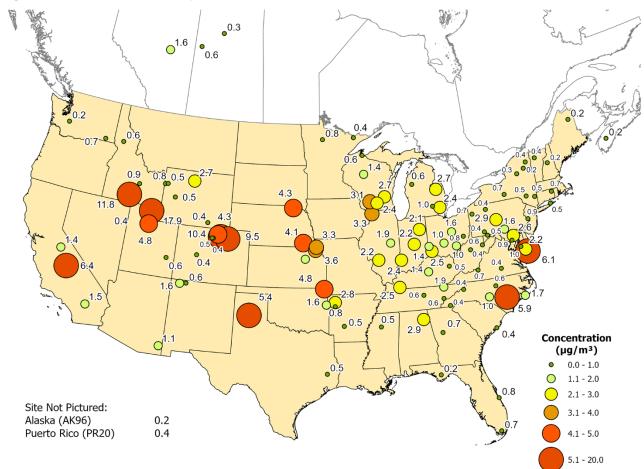
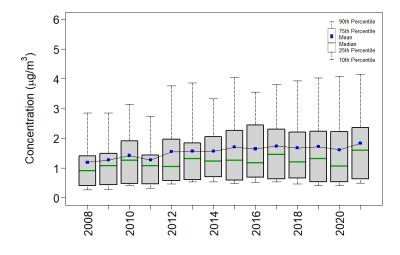


Figure 13 Annual Mean NH₃ Concentrations at AMoN sites for 2021





Continuous Trace-level NO_y and Filter Pack Total Nitrate Concentrations

NO_y is defined as NO_x (NO + nitrogen dioxide) plus NO_z (HNO₃, nitrous acid, peroxyacetyl nitrate, peroxyproyl nitrate, other organic nitrates, and nitrite). It consists of reactive gases that are considered precursors of O₃ and PM_{2.5}. Because HNO₃ and particulate NO₃ are measured as components of NO_y, NO_y concentrations should always be higher than total NO₃ levels (i.e., the ratio of NO_y to total NO₃ should always be greater than 1.0). A comparison of weekly mean continuous NO_y concentrations with filter pack total NO₃ levels at Mammoth Cave National Park, KY (MAC426); Huntington Wildlife Forest, NY (HWF187); and Pinedale, WY (PND165) for 2021 was used to evaluate the 2021 measurements (Figures 15 through 17). The scales for NO_y and total NO₃ levels are different in the three figures. The NO_y concentrations were consistently higher than the total NO₃ levels, as expected.

The results are similar for all seven sites operating continuous NO_y analyzers. The weekly total NO₃ concentrations, the average weekly NO_y levels, and their ratios are listed in Table 3. These were calculated as the average of all valid weekly filter pack concentrations, the average of mean NO_y values matching the run time of the weekly filter packs, and the average of the ratios calculated for each week.



Salamonie Reservoir (SAL133) - Wabash, IN

Figure 15 Comparison of MAC426, KY Weekly Mean Continuous Trace-level NO_y and Filter Pack Total NO₃ Concentrations

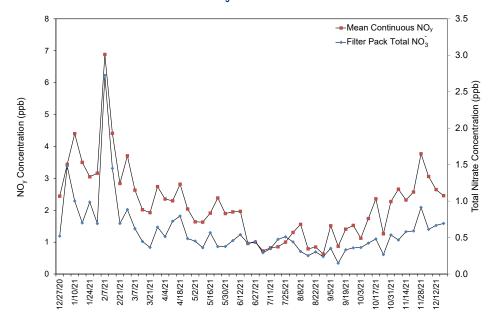


Figure 16 Comparison of HWF187, NY Weekly Mean Continuous Trace-level NO_y and Filter Pack Total NO₃ Concentrations

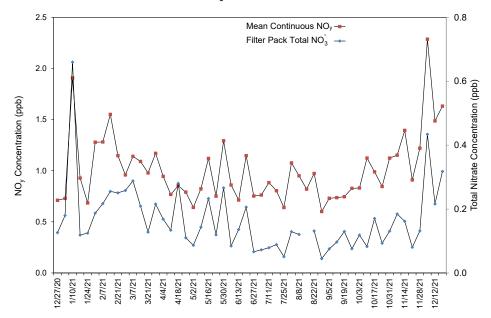


Figure 17 Comparison of PND165, WY Weekly Mean Continuous Trace-level NO_y and Filter Pack Total NO₃ Concentrations

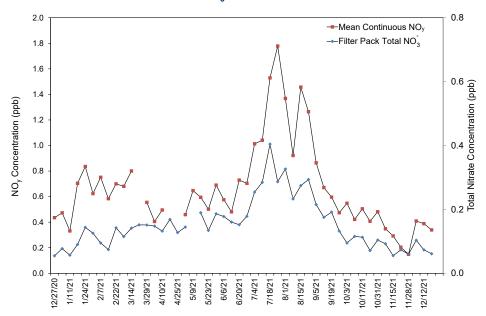


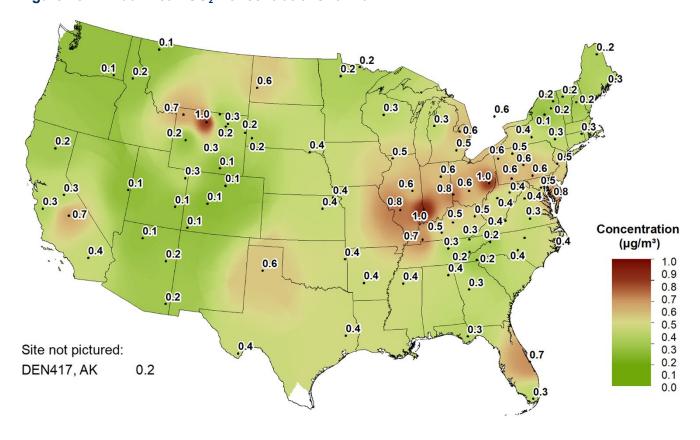
Table 3 Summary of Total NO₃ and NO_y Measurements for 2021

Site Location	Elevation (m)	Total NO ₃ (ppb)	NO _y (ppb)	Ratio
BVL130, IL	213	0.89	4.03	4.87
MAC426, KY	243	0.59	2.19	3.89
HWF187, NY	497	0.17	1.01	7.24
GRS420, TN	793	0.38	1.52	4.29
PNF126, NC	1216	0.29	0.89	3.09
PND165, WY	2386	0.15	0.66	4.63
ROM206, CO	2742	0.21	1.08	5.62

Sulfur Atmospheric Concentrations

Annual mean concentrations of SO_2 and SO_4^{2-} for 2021 are presented in Figures 18 and 20, respectively, in this section. Trends in annual mean concentrations are also shown in Figures 19 and 21. Sites that did not meet the completeness criterion appear as a dot with no value on the various concentration maps.

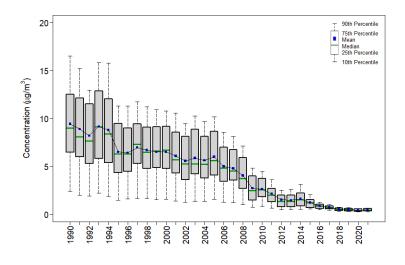
Figure 18 Annual Mean SO₂ Concentrations for 2021



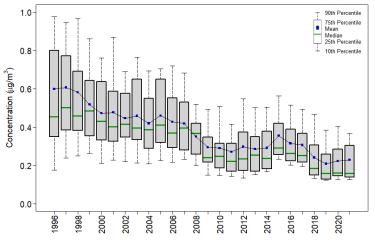


Nez Perce Tribe (NPT006) - Harrisburg, ID

Figure 19 Trends in Annual Mean SO₂ Concentrations



Eastern Reference Sites



Western Reference Sites

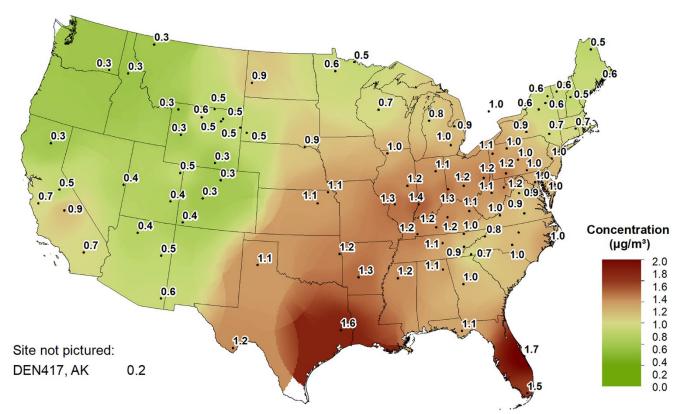
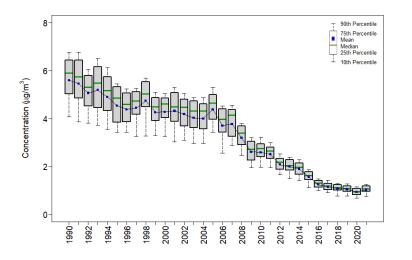


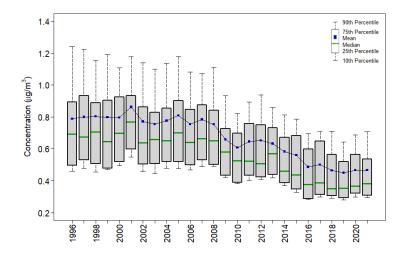
Figure 20 Annual Mean SO₄² Concentrations for 2021

Figure 21 Trends in Annual Mean SO₄² Concentrations (1 of 2)



Eastern Reference Sites

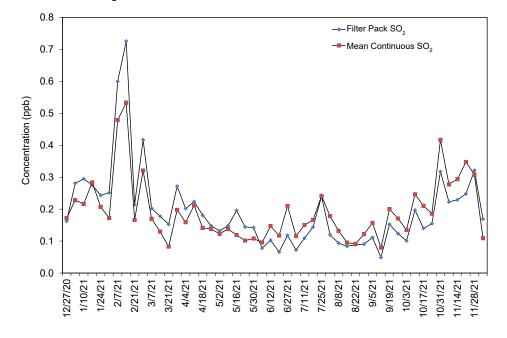
Figure 21 Trends in Annual Mean SO₄² Concentrations (2 of 2)



Western Reference Sites

Figure 22 provides a diagram that compares weekly filter pack SO_2 concentrations with continuous trace-level gas data measured at MAC426, KY. The continuous analyzer provides a gross check of the filter pack measurements. The continuous analyzer data can also be used to produce diurnal profiles of hourly SO_2 concentrations and detect the presence of SO_2 plumes.

Figure 22 Comparison of MAC426 Weekly Mean Continuous and Filter Pack SO₂ Concentrations



Wildfire Impacts to Ozone and Biomass Concentrations at Rocky Mountain National Park

The co-located CASTNET monitoring sites at ROM406/206 (Figure 1) are in the Northern Front Range in Larimer County, Colorado. Ozone and meteorological measurements began at ROM406 in 1987 as part of the NPS Gaseous Pollutant Monitoring Program and were adopted into CASTNET in 1994. Measurements at ROM206, the EPA-operated site of the pair, began in 2011. Despite being within a Class I protected area, ROM406/206 is frequently impacted by local and regional air pollution. ROM406/206 measured an O₃ design value of 71 ppb for the three-year period 2019–2021, which is above the 2015 O₃ NAAQS.

In Colorado, two of the largest wildfire seasons on record occurred in 2018 and 2020, including five wildfires greater than 100,000 acres. The largest wildfire in Colorado history, the Cameron Peak fire, began in August 2020 and ended in December 2020. It consumed 209,000 acres and destroyed 469 structures The fire covered sections of Rocky Mountain National Park. Seventeen O₃ exceedances were observed at ROM406/206 in 2018 and 2020, with over half of them likely enhanced by wildfires. This section discusses the relationship between O₃ concentrations, filter pack measurements, tracelevel gas measurements, and satellite-based wildfire smoke data.

Wildfires

Wildfires generate particulate matter and O_3 precursors that may influence tropospheric O_3 levels depending on meteorological conditions and smoke density. Dense wildfire smoke may block ultraviolet light and subsequently reduce O_3 photochemical production. Wildfires emit cellulose and hemicellulose compounds including levoglucosan, which is a common wildfire indicator due to its high concentrations (e.g., 40 to 1,200 milligrams per kilogram) and relative stability within the atmosphere (Simoneit *et al.*, 1999; Simoneit, 2002).

Measurements

In addition to standard CASTNET measurements, other data were collected at ROM406/206 from 2018 through 2020 to help understand how wildfires impact air quality:

- Hourly NO and NO_v were measured using a Teledyne API T200U NO_v analyzer.
- Daily concentrations of biomass compounds, including levoglucosan and other organic carbohydrates, sugars, sugar alcohols, and base cations, were measured using a URG 47-millimeter nylon filter by Colorado State University (CSU).
- The URG nylon filters were analyzed at the CSU laboratory using high-performance anion exchange chromatography with pulsed amperometric detection; for more on this analysis, see Sullivan *et al.* (2011a; 2011b).
- In June 2020, the URG sample period was changed from daily to every three days.

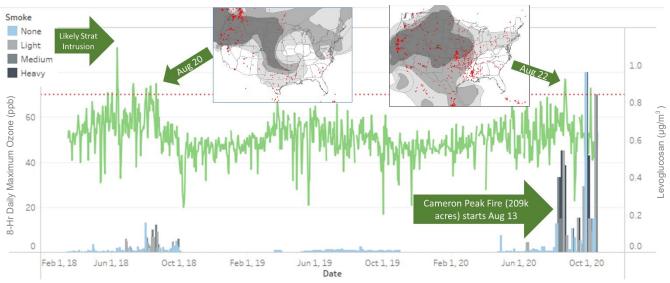
Methods

Wildfire smoke data from NOAA's Hazard Mapping System (HMS) were used to characterize the impact of smoke on air quality at ROM406/206 for 2018 and 2020. HMS wildfire data from 2019 were not included due to limited wildfire activity. The HMS wildfire data consisted of 2-km horizontal resolution measurements from the two Geostationary Operational Environmental Satellites (GOES-16 [East] and GOES-17 [West]). The smoke density at the sites was classified using the three HMS smoke density levels of light, medium, and heavy, which approximate smoke concentrations of 0 to 10 micrograms per cubic meter (µg/m³) of air, 10 to 21 µg/m³, and 21 to 32 µg/m³, respectively (McNamara *et al.*, 2004; NOAA, 2019). DM8A O₃ concentrations were calculated and compared with daily biomass (as levoglucosan) concentrations, NOAA HMS smoke density data, daily averages of trace-level NO and NO₂ measurements, and weekly CASTNET filter pack concentrations. Daily NOAA HMS smoke densities for ROM406/206 were mapped/intersected using ArcGIS ArcMap v10.8.1. The densest smoke level was used when multiple smoke densities occurred on the same day.

Preliminary Results

Twelve O_3 exceedances were observed in 2018 and five exceedances occurred in 2020 (Figure 23). Five of the 12 O_3 exceedances in 2018 had levoglucosan levels above the annual 75th percentile (0.018 μ g/m³), and 6 of these 12 O_3 exceedances showed NOAA HMS smoke events (any density). Four of the five 2020 O_3 exceedances had levoglucosan levels above the annual 75th percentile (0.127 μ g/m³), and all five exceedances indicated NOAA HMS smoke events. On June 11, 2018, the DM8A O_3 value of 91 ppb was flagged as an informational stratospheric O_3 intrusion ("IO") by request of the Colorado Department of Public Health and Environment (Harshfield, 2019). Continued work will include an assessment of wildfire plume impacts on air quality across the network using available wildfire data resources.





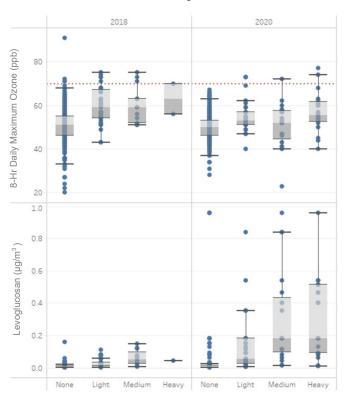


Figure 24 DM8A O₃ and Levoglucosan Concentrations as a Function of NOAA HMS Smoke Density

As shown in Figure 24, the median DM8A O₃ value in 2018 was 51 ppb without NOAA HMS smoke and was 59 ppb, 59 ppb, and 63 ppb under light, medium, and heavy smoke density, respectively, resulting in a smoke-enhancement of 8 to 12 ppb. In 2020, the median DM8A O₃ value was 50 ppb without smoke, but was 53 ppb, 52 ppb, and 55 ppb under light, medium, and heavy smoke density, respectively, a smoke-enhancement of 2 to 5 ppb. The 2020 median levoglucosan concentrations ranged from 3.1 to 6.3 times higher than values from 2018, and both years had positive correlations with smoke density.



Rocky Mountain National Park (ROM406/206) - Larimer, CO (co-located)

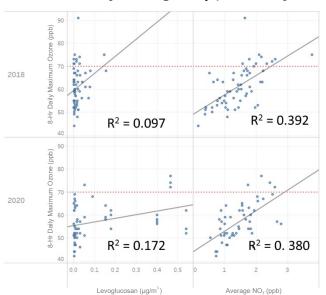


Figure 25 Linear Regression Plots of 8-hour Daily Maximum O₃ versus Levoglucosan and Daily Average NO₇ (June, July, and August 2018 and 2020)

Figure 25 displays linear regressions of 2018 and 2020 DM8A O_3 concentrations compared against levoglucosan and daily average NO_y concentrations in summer (i.e., June, July, and August) when precipitation was less than 0.5 cm. There is a faint positive trend between O_3 and levoglucosan and a slight positive trend between O_3 and NO_y . Other preliminary results include:

- Ozone, levoglucosan, and wildfire smoke events exhibit a complex relationship.
- Ozone exceedances in 2018 and 2020 occurred when the NOAA HMS smoke density data showed wildfire smoke impacting the site, suggesting wildfire smoke is influencing O₃ exceedances at Rocky Mountain National Park.
- Despite a weak linear relationship between O₃ and levoglucosan levels (Figure 25), both analytes correlate with increasing values as a function of NOAA HMS smoke density.
- Daily average NO_y showed the strongest positive relationship with O₃ of the analytes measured.

Recommended Additional Work

- Additional work is needed to relate the other URG biomass compound signatures (e.g., ratios
 of carbohydrates to sugars) as they relate to individual wildfires.
- Additional work beyond smoke density is needed to understand the effects that wildfire size, distance, and duration play on the magnitude of smoke impact to O₃ enhancement.
- Exploration of methods for getting biomass concentrations from the current CASTNET filter pack measurements is needed.

Atmospheric Deposition

Gaseous and particulate nitrogen, sulfur, and other trace pollutants are deposited to the environment through dry and wet atmospheric processes. A primary goal of CASTNET is to estimate the dry deposition of pollutants from the atmosphere to sensitive ecosystems. The NADP TDep measurement-model technique (Schwede and Lear, 2014) combines CASTNET monitoring data with output from the Community Multiscale Air Quality (CMAQ) modeling system (Byun and Schere, 2006) to estimate dry deposition. The TDep method uses measured air concentration data from CASTNET when available and CMAQ output in areas where monitoring data are not available. CMAQ also provides modeled data for species that are not routinely measured. TDep and its recent updates are discussed on the TDep web page (https://nadp.slh.wisc.edu/committees/tdep/). See also the NADP Fact Sheet, "Hybrid Approach to Mapping Total Deposition," https://nadp.slh.wisc.edu/wp-content/uploads/2021/04/tdepsheet 2021.pdf.

For wet deposition, the TDep method combines precipitation amounts measured at NADP/NTN sites with a Parameter-elevation Regressions on Independent Slopes Model (PRISM) surface of precipitation data to develop a continuous grid of precipitation data. PRISM uses terrain elevation, slope, and aspect and climatic measurements to estimate precipitation amounts. Pollutant concentrations in precipitation are measured by NADP/NTN, extrapolated into a grid using an inverse distance weighting algorithm, and then multiplied by the combined precipitation grid to generate estimates of wet deposition. For 2021, dry and wet deposition rates were added to obtain estimates of total deposition of nitrogen, sulfur, base cations, and Cl⁻, which are presented on the maps in this chapter as kilograms per hectare per year (kg ha⁻¹ yr⁻¹) unless otherwise noted. The magnitude of the deposition fluxes is illustrated by the shading in the figure legends.



Santee Sioux (SAN189) - Knox, NE

Nitrogen Deposition

A map of the dry deposition fluxes of nitrogen (N) estimated for 2021 by the TDep method is shown in Figure 26. A map of total deposition of N estimated for 2021 is given in Figure 27. The percentage of total deposition of N due to dry deposition is shown in Figure 28. A map of TDep estimates of total deposition of reduced N species for 2021 is given in Figure 29. Figure 30 gives the percentage of total deposition of N from reduced N. Dry deposition of N species that are not routinely measured but are modeled by CMAQ and used by TDep is shown in the map in Figure 31, and its percentage of total deposition is shown in Figure 32. The map in Figure 33 shows estimates of annual gross dry deposition of NH₃ as N for 2021.

Sulfur Deposition

Figure 34 shows a map of TDep estimates of total deposition of sulfur (S) for 2021. The percentage of total deposition of S due to dry deposition is shown in Figure 35.

Deposition of Base Cations and Chloride

A map of 2021 estimated total deposition of the base cations Ca²⁺, K⁺, Mg²⁺, and Na⁺ is provided in Figure 36. Figure 37 provides 2021 estimated rates of deposition of total Cl⁻.

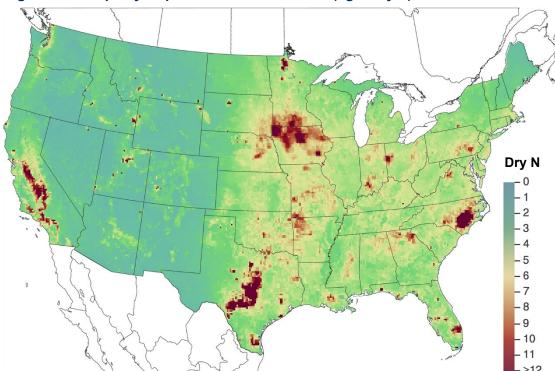


Figure 26 TDep Dry Deposition Estimates of N (kg ha⁻¹ yr⁻¹) for 2021

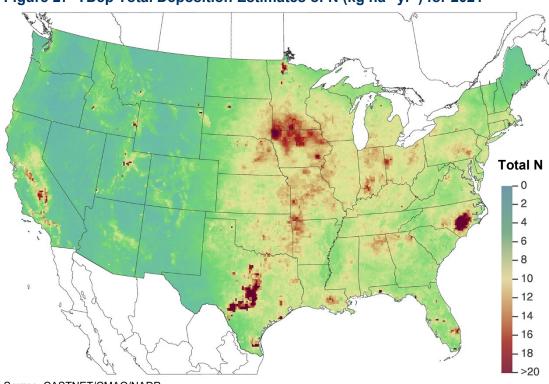


Figure 27 TDep Total Deposition Estimates of N (kg ha⁻¹ yr⁻¹) for 2021

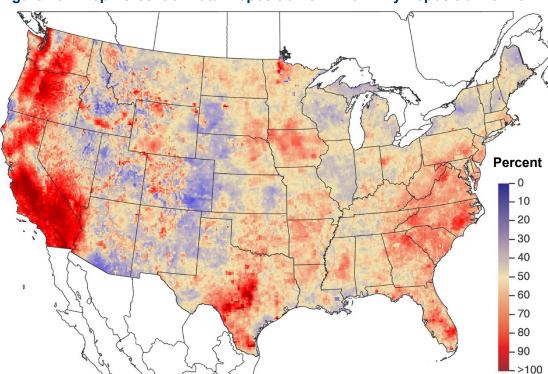


Figure 28 TDep Percent of Total Deposition of N from Dry Deposition for 2021

Total Reduced N

-0
-1
-2
-3
-4
-5
-6
-7
-8
-9
>>10

Figure 29 TDep Total Deposition Estimates of Reduced N Species (kg ha⁻¹ yr⁻¹) for 2021

Source: CASTNET/CMAQ/NADP

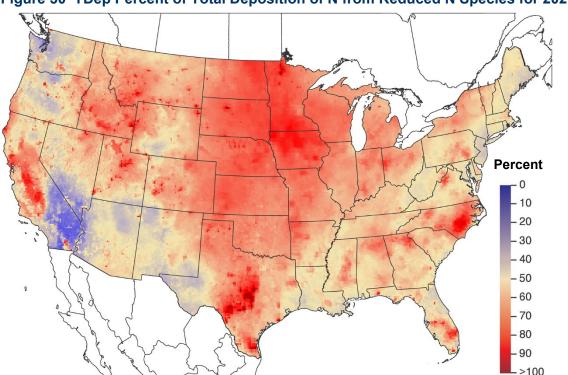


Figure 30 TDep Percent of Total Deposition of N from Reduced N Species for 2021

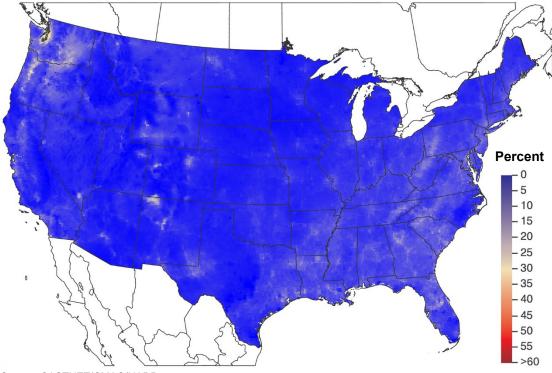
35

Dry N

-0.0
-0.5
-1.0
-1.5
-2.0
-2.5
-3.0
-3.5
-4.0
-4.5
>>5.0

Figure 31 TDep Dry Deposition Estimates of Unmonitored N Species (kg ha⁻¹ yr⁻¹) for 2021

Figure 32 TDep Percent of Total Deposition of N from Dry Deposition of Unmonitored Species for 2021



Dry N

-2.0

-3.0

-4.0

-5.0

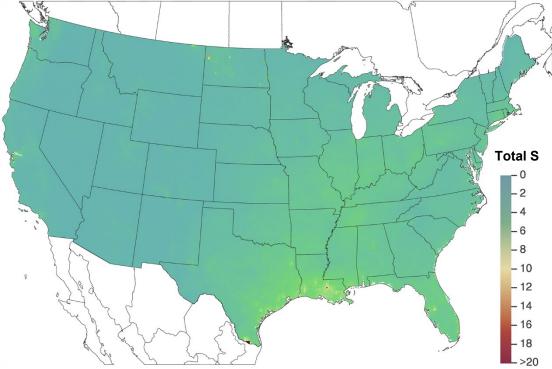
-6.0

-7.0

>8.0

Figure 33 TDep Gross Dry Deposition Estimates of NH_3 as N (kg ha⁻¹ yr⁻¹) for 2021





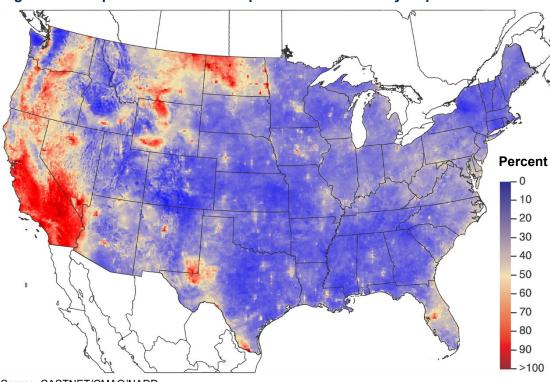
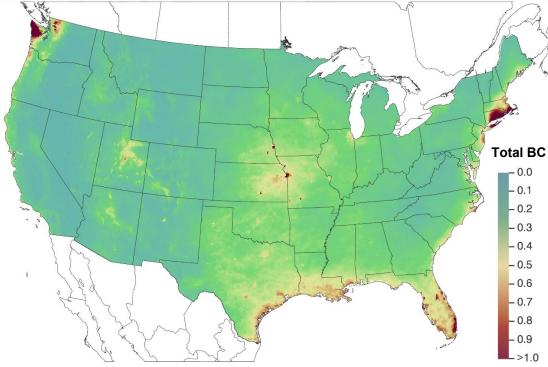


Figure 35 TDep Percent of Total Deposition of S from Dry Deposition for 2021





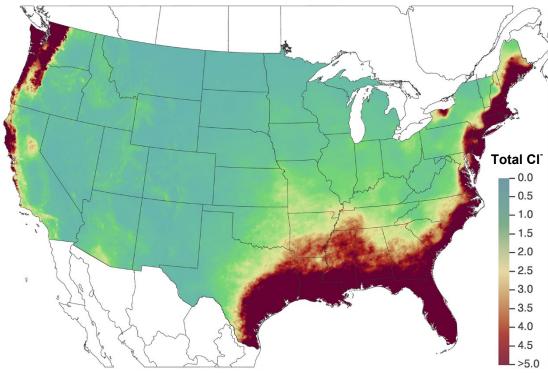


Figure 37 TDep Total Deposition Estimates of Cl (kg ha⁻¹ yr⁻¹) for 2021

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